



Supporting Online Material for

Rebuilding Global Fisheries

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Methods and Data Sources

In the following, we describe in detail the methods and data sources supporting our analyses of (i) ecosystem models, (ii) stock assessment data, (iii) trawl survey data, (iv) catch data, and (v) fishing access data. Data sources are described in Tables S1-S3 and supporting data are provided in Tables S4-S7 and Figures S1-S6.

Ecosystem models

We compiled 37 published ecosystem models from 31 systems (Table S1, mapped in Fig. 1). These models were either ECOPATH with ECOSIM ($S1$, $S2$) or ATLANTIS ($S3$) models. They were selected because they were publicly available, have been properly documented and quality controlled, and covered all systems that we examined empirically using the stock assessment and trawl survey data. For each ecosystem model, we did the following:

1) For each fished group in the model the exploitation rate u (defined as the proportion of biomass that is removed per year, i.e. $u_t = C_t / B_t$ where C is the catch (or ‘yield’) and B is the available biomass in year t) was incrementally increased and decreased, holding u constant for all other groups in the model, but allowing for full dynamic responses due to trophic interactions and direct fisheries extractions. This step produced an estimate of the exploitation rate that produced maximum sustainable yield u_{MSY} for each species.

2) The u for the fished groups in the model were set to $k \times u_{MSY}$ from step 1. Long-term runs (1000 years) were then run with k incrementing from $k = 0$ to $k = 20$. 1000 years was selected because EwE models are effectively equilibrium models and the simulations had returned to a stable state by 1000 years.

3) While predation was allowed to vary during step 1, in reality u would also vary across species through time (with changing targeting and gear use). Also, given ecosystem interactions can be complex it was felt that it would be beneficial if additional runs were done to try and further fill out the phase space of possible outcomes when fishing across an entire ecosystem.

Consequently, to complement the MSY based exploration, a set of fisheries policy searches was

performed. The objective function used in the search combines economical, ecological and potentially social terms. The set of searches incremented from the policy with the economic contribution to the objective function set to 1 (effectively maximizing catch from the system as there was little actual economic information included in the model formulations) and the ecological contribution (ecosystem structure and mandated rebuilding) set to 0 through to the opposite weightings (economic weighting = 0, ecological = 1). Levels of mandated rebuilding from base ECOPATH levels were set based on how the base ECOPATH model biomasses compared with 40% of the unfished biomass levels (taken from a long-term run of the ECOSIM model with all u set to 0). Values relative to the unfished run were used, because many of the ECOPATH models included heavily exploited groups (that were in an already depleted state and so simply setting rebuilding to 0.4 of ECOPATH values would be inappropriate). Ideally social (employment) considerations would have also been included in this policy search. Unfortunately, employment information was not readily available for many of the modeled fleets and the same fleet resolution was not available in all models, so social contributions to the objective function were not considered.

4) The results from the different analyses per system (i.e. the results of steps 2 and 3) were then combined to produce aggregate plots of catch, available biomass, size, and the number of groups that have dropped below 10% of their unfished levels (defined as ‘collapsed’) against the system-level exploitation rate (calculated as the catch / total available biomass).

5) Using the catch-exploitation rate plot for each system, the point of MMSY was defined as the peak of the plot, with the bands of uncertainty defined as the exploitation rates producing 90% of that peak catch. Similarly the conservation reference point $u_{conserve}$ (10% of stocks collapsed) and associated uncertainty band was read off the model plots, with the band defined as the u giving 9-11% of the stocks collapsed. Where alternative parameterizations for the model existed (e.g. for the Benguela and SE Australia) all steps 1-4 and calculation of MMSY and conservation reference points were repeated and in all cases the resulting values from the alternative parameterizations fell within the uncertainty bands defined from the original plots. Where there was any difference in the uncertainty bands produced by the alternative parameterizations these were combined to give the final uncertainty band used in Fig 3A. In addition, in some cases it was necessary to combine the results of multiple models to produce results at an LME or regional scale for Fig 3A, such as the Newfoundland-Labrador Shelf LME which includes the Northern Gulf St Lawrence and Grand Banks models, and the Celtic-Biscay LME which includes the Bay of Biscay, Irish Sea, and Western English Channel models. In those cases the final bands were created using an area-based weighted average (so a model covering a smaller section of the overall area contributed less to the average).

6) An overall plot (Fig. S1) was created by averaging the results in step 4 across all ecosystems, the confidence bounds in the plot mark one standard error from the mean.

Comparing ecosystem and single-species MSY

In some cases (e.g. Georgia Strait, Southeast Australian shelf) the sum of the predicted single-species MSY did approximate (within 2%) the system-level catches ($MMSY$). (S4) found this from some systems, even when at an individual level the realized catches of species within those systems could be strongly different to the predicted single-species MSY and often at the expense

of top predators. However, it would be inappropriate to use this result as a “rule of thumb” to predict what $MMSY$ would be for a system, as the sum of single-species MSY is generally a poor predictor of multi-species MSY . This is because it’s hard to say *a priori* whether compensatory or compensatory responses to fishing will occur. These responses will lead to divergence between yields at the system level and those predicted by single-species assessments as seen in (S5). The sum of predicted single-species MSY differed from system-level $MMSY$ by more than 20% in 42% of the systems and by more than 50% in 18% of the systems.

Across the modeled systems there were examples of systems for which the sum of single-species MSY exceeded $MMSY$ (e.g. Great Barrier Reef, North Sea); however there were also cases (e.g. Central North Pacific, Bay of Biscay) where the sum of single-species MSY was less than $MMSY$ (but at a significant cost to top and even medium level predators). Looking at system type (inshore versus shelf or open ocean; temperate versus tropical), there is no consistent pattern regarding how $MMSY$ at a system level will compare with the sum of single-species MSY . In contrast, patterns of response are clearer at an individual group level. For individual groups, across all systems, it was found that $MMSY$ is never significantly greater than MSY at the higher trophic levels, whereas this is often the case at lower trophic levels. In the majority (61-71%) of cases the catch, when the system was fished at u_{MMSY} , was greater than the catch predicted from single-species models for groups of trophic level 1 or 2. For trophic level 3, there was an even split in the number of cases in which the catch of a group under u_{MMSY} exceeded the expected MSY and vice versa (i.e. $MSY > MMSY$ for that group in 50% of cases). For the highest trophic levels (4+) in the majority of cases (66-84%) the group-level catch under u_{MMSY} is much less than MSY . For those cases in which $MMSY$ is not less than MSY then they are effectively identical (differing by less than 4%).

Creating the exploitation-rate plots

To give insight into the history of the exploitation of a range of systems from around the world, reference points were taken from the ecosystem modeling analysis (see step 5 above) and compared with the history of overall exploitation rates observed (calculated from catch and biomass for assessed species) in those same systems (Fig. 3A, Table S4, see stock assessment data). The reference points were (i) u_{MMSY} : the band of exploitation rates that produced the maximum system-level catch (uncertainty associated with the models and also the shape of the curves meant there was no single peak exploitation rate, but rather a band of potential rates) and (ii) $u_{conserve}$: the band of exploitation rates where 10% of the groups in the model fell below 10% of their unfished biomass levels ($u = 0$ for all system components). This latter reference point was chosen as a hypothetical conservation objective; as there is no easily defined conservation equivalent of an MSY concept, the use of other levels of extirpation could be substituted.

In Fig. 3A, the ratio of biomass B to B_{MSY} was calculated for each of these systems, from corresponding units, the ratio of spawning biomass to the spawning biomass that would be present in a system producing MSY , but sometimes total biomass for both. Where available, these ratios were obtained from the stock assessments, otherwise B_{MSY} was calculated using the surplus production model approached outlined below.

Stock assessment data

We gathered time series of recorded catch, model-estimated biomass and fishing mortality rates, and reference points (B_{MSY} and u_{MSY} , the biomass and exploitation rate, respectively, that result in maximum sustainable yield) from published stock assessments of exploited marine fish and invertebrate stocks and from personal communications with stock assessment scientists (see Table S2). We attempted to gather the most recent available assessments for stocks around the world in which a population model was applied to estimate a time series of biomass and exploitation rates. While we compiled the majority of assessments we have identified, the data set is not complete and is evolving. Many regions are under-represented either because of analytical uncertainties associated with recent attempts to assess the stocks or because assessments for previously over-exploited stocks are highly uncertain or are no longer conducted. A prime example is that of northern cod off Newfoundland and Labrador, a stock estimated to have declined 99% between 1962 and 1992, for which the offshore segment of the stock has not been assessed since shortly after it collapsed in 1992.

Where assessment estimates of B_{MSY} and u_{MSY} (or the instantaneous fishing mortality rate F_{MSY} which was provided for some stocks) were available ($N = 41$ stocks for u_{MSY} or F_{MSY} and $N = 54$ stocks for B_{MSY}), these were used to determine stock status (Fig. 3B and Fig. 4M-P). Where these reference points were unavailable, we fit a surplus-production model to time series of annual total biomass B_t and total catch or landings (where catch was unavailable) C_t from the assessments. The surplus-production model was only applied where ≥ 20 years of catch or landings and biomass data were available. One hundred and sixty-six of the 239 stock assessments that we gathered either had B_{MSY} and u_{MSY} reference points provided or had ≥ 20 years of catch or landings and biomass data.

Overall biomass trends computed from stock assessments (Fig. 4A-D) are provided in Table S5.

Surplus-production model

Surplus production in year t , P_t , a commonly-used measure of stock productivity has been used previously (S6, S7, S8) where surplus production in year t , P_t , can be calculated as:

$$(1) \quad P_t = B_{t+1} - B_t + C_t$$

where B_t is the biomass at time t and C_t is the catch at time t .

We fit a Schaefer surplus-production model, which is based on a logistic model of population growth (S9) to the catch and biomass time series data. The predicted surplus production in each year in the Schaefer model is given by:

$$(2) \quad \hat{P}_t = \frac{4mB_t}{K} - 4m\left(\frac{B_t}{K}\right)^2$$

where m is the maximum sustainable yield and K is the carrying capacity or equilibrium biomass in the absence of fishing.

We estimated the model parameters (m and K) using maximum likelihood in AD Model Builder (<http://admb-project.org>) assuming that the residuals ($\varepsilon_i = P_i - \hat{P}_i$) were normally distributed.

For the Schaefer model, B_{MSY} is simply $0.5K$, and the harvest rate that results in maximum sustainable yield, u_{MSY} , is m/B_{MSY} . Carrying capacity was constrained to be less than twice the maximum observed biomass. Thirty-eight percent of stocks were affected by this constraint. For five Eastern Bering Sea crab stocks, only B_{MSY} reference points were available, not u_{MSY} reference points and surplus production model fits could not be obtained, therefore the total number of stocks for which we were able to obtain estimates of $B_{current}/B_{MSY}$ and $u_{current}/u_{MSY}$ was 160 (Fig. 3B).

We compared the surplus production model estimates of $B_{current}/B_{MSY}$ and $u_{current}/u_{MSY}$ to the value of these ratios obtained from the assessments for all stocks which had assessment-based reference points. After replacing values of these ratios that were greater than 2 with a value of 2 (as was done in Fig. 3B), the Pearson correlation between the harvest rate ratios from the surplus production model and the harvest rate ratios from the assessments was $r = 0.62$. The equivalent correlation for the biomass ratios was $r = 0.65$. All stocks used in this analysis and their estimated $B_{current}/B_{MSY}$ and $u_{current}/u_{MSY}$ are shown in Table S2.

Caveats on MSY-related calculations

Despite the fact that the concept of *MSY* has been prominent in the fisheries science and management literature for about 5 decades, it is not uniformly defined or estimated. Numeric values of B_{MSY} and u_{MSY} are dependent on the vulnerability of different ages or sizes of fish to fishing gear; whether B_{MSY} is defined in terms of spawning biomass, available (exploitable) biomass, or total biomass; the harvest strategy used or assumed (e.g. a constant catch strategy or a constant fishing mortality strategy); and the model or method used for estimation. The Schaefer surplus-production model used here when the management agency did not provide its own estimates of B_{MSY} assumes a symmetric relationship between sustainable yield and biomass (with B_{MSY} being half of the carrying capacity) whereas the more common result is that B_{MSY} is less than half of carrying capacity, usually in the range 25-40%. If exploitation can be delayed until several years after the onset of maturity, then this range will be even lower.

Use of asymmetric models will undoubtedly give different estimates of the MSY related ratios. This coupled with our decision to use the estimates provided with assessments in preference means that the ratio estimates may not be fully consistent. However, we believe that our analysis represents a first attempt to characterize the status of a large number of fisheries worldwide in terms of both $B_{current}/B_{MSY}$ and $u_{current}/u_{MSY}$. Detailed analysis of individual stocks will undoubtedly come to different conclusions if other models or data are used.

Finally, catch time series presented in the assessments should match time series of exploitation estimates derived from the instantaneous fishing mortality rates. In cases where these differ, for example when unaccounted discards or misreporting lead to commercial data being omitted or down-weighted in the assessment, the ratios of exploitation rates to the reference exploitation rate will also differ. In a preliminary investigation of ICES stocks, for example, differences between ratios based on exploitation estimates from catches and mortality rates were typically small, with the important exceptions of West of Scotland cod and haddock, and Irish Sea cod.

For these stocks the exploitation ratios were higher when based on converted fishing mortality rates than when using catches or landings. The differences follow from our decision to use a consistent method for describing exploitation status, but they do not affect our main conclusions at regional and global scales. However, the differences do highlight the importance of consulting the original assessments and assessment scientists when seeking information on the status of individual stocks.

Methods to obtain trends in relative biomass and exploitation rates

The biomass trends in Fig. 3A for each Large Marine Ecosystem were obtained by taking the geometric mean of the B/B_{MSY} ratios from assessed stocks in each year (Table S4). The geometric mean was preferred to an arithmetic mean since these are averages of ratios. The exploitation rate trends were obtained by calculating the ratio of total catch to total assessment biomass in each year. Three species were excluded from the trajectories of biomass and exploitation rate in Fig. 3A, these species were pelagic species with dominant catches but these catches fell mostly outside the Large Marine Ecosystems examined. The excluded species were: Pacific hake (*Merluccius productus*) for the California Current, Atlantic menhaden (*Brevoortia tyrannus*) in the Northeast U.S. Shelf, and blue whiting (*Micromesistius poutassou*) in the Celtic-Biscay Shelf.

Methods to calculate trends in collapsed taxa

To calculate trends in the proportion of total stocks collapsed (Fig. 4M-P, Table S7), we compared time series of biomass B to B_{MSY} . The proportion of collapsed stocks in any year is simply the number of collapsed stocks divided by the total number of stocks for which an estimate of biomass was available in that year. A stock was defined as collapsed in any year if the biomass in that year was less than 20% of B_{MSY} . For a population growing according to the logistic growth function, this is equivalent to 10% of carrying capacity.

Trawl survey data

We compiled data from 20 long-term research trawl surveys (Fig. S2, Table S3), from a variety of regions around the world, but dominated by the Northern Hemisphere ($n = 17$) and especially the Northwest Atlantic ($n = 10$). The surveys each spanned at least 18 years from earliest to the latest and contained at least ten annual surveys. Surveys were typically obtained directly from the agency responsible for the surveys, but also from published sources.

Each taxon in each survey was allocated to a category: invertebrate, pelagic fish and demersal fish. The demersal fish were further subdivided into small (≤ 30 cm), medium (30-90 cm) and large (≥ 90 cm) (as plotted in Fig. S2) categories based on the maximum length (L_{max}) recorded in the online database FishBase (S10), or SeaLifeBase (S11), where available. Where L_{max} was missing for a particular species, these were assumed to be the average for that genus or where that was not possible, from family or higher-level taxon. In the rare instances where all of these methods failed in obtaining L_{max} values ($< 1\%$ of the total), they were obtained from a variety of grey literature and internet sites. Allocation to demersal or pelagic were based on the habitat categories in FishBase with categories of bathydemersal, benthopelagic, and reef-associated all

assumed to be “demersal”, and categories of bathypelagic, pelagic-neritic and pelagic-oceanic all assumed to be “pelagic”. Although the default option for benthopelagic species was to assume they were demersal, the following species of obviously pelagic nature (Atlantic herring *Clupea harengus*, American butterfish *Peprilus triacanthus*, black mackerel *Scombrobrax heterolepis*, oxeye herring *Megalops cyprinoides*, red tailed round scad *Decapterus russelli*, and jack mackerel *Trachurus declivis*) were categorized as pelagic.

The survey data typically came from gears designed to adequately sample medium to large demersal species. While recognizing this limitation for interpreting trends in other categories, catch trends were included if deemed a reliable index by the agency that supplied the data.

Invertebrates

Invertebrates for a given survey were classified as “not recorded” if a data provider supplied finfish data only (Table S3). This was the case for: St Pierre Bank, Newfoundland; Southern Grand Banks, Newfoundland; Southern Gulf of St. Lawrence; Northern Gulf of St. Lawrence; Celtic Sea; and the North Sea IBTS surveys. In the Scotian Shelf survey data invertebrates are represented by one species (*Illex illecebrosus*). No other invertebrate species were recorded in the data provided. A separate invertebrate survey dataset for this region is only available from the year 2000, when invertebrates were consistently recorded. Given the short length of this series, it was not included in the analyses.

Demersal ≤ 30 cm

When the data providers deemed catch trends for small demersals unreliable, “not recorded” values were assigned. This was the case for: St Pierre Bank, Newfoundland; Northern Gulf of St Lawrence; and Eastern Bering Sea surveys. In the Eastern Bering Sea survey grouped categories like “Sculpins” might have contained some component species that were small demersals, but the groups as a whole were all >30 cm. True zeroes were provided for small demersals in surveys where the biomass of small demersal species contributed to less than 2% of the total biomass sampled per year. This was the case for: URI Whale Rock and URI Fox Island surveys.

Pelagic

Similar to small demersals, pelagic species sampling is limited by the demersal nature of the sampling gear. Here again, if trends in pelagic species were deemed a reliable index by the agency that sampled them, they were provided and are included in the analyses. Pelagic species were assigned “not recorded” values because they were not provided in the following surveys: St Pierre Bank, Newfoundland and Northern Gulf of St Lawrence.

Methods to obtain overall trends in survey biomass

Overall survey biomass trends (Fig. 4E) were based on 19 of the surveys in Table S3 and are shown in Table S6. We did not include the South Georgia series in the analysis because only five commercially important species were reported in the source document and these were not representative of the ecosystem as a whole. We also excluded years where fewer than four surveys were represented in the data set.

Data standardization

The survey biomass index (biomass, biomass per unit area/tow-time, depending on the survey) of taxon i in survey s in year t was given by $B_{i,s,t}$. The i taxa (usually identified to species but occasionally only identified to genus or family) were grouped into one of $c = 5$ categories as described above (Pelagic fish, Invertebrate, Demersal fish ≤ 30 cm, Demersal fish 30-90 cm, Demersal fish ≥ 90 cm). The biomass of category c in survey s in year t was given by

$$(3) \quad B_{c,s,t} = \sum_{i=1}^{n_{c,s}} B_{i,s,t}$$

where $n_{c,s}$ is the number of taxa in category c in survey s .

As the data come in a variety of different units, it was necessary to standardize before analyzing the trends. A standardization method was required that maintained the strength of the category within a survey but also allowed the categories to be combined across surveys. Assuming that the biomass indices from a given survey are lognormally distributed within a year, the index was log transformed:

$$(4) \quad b_{c,s,t} = \ln(B_{c,s,t} + 1)$$

The data were standardized by subtracting the survey mean (on the log scale), that is:

$$(5) \quad \delta_{c,s,t} = b_{c,s,t} - \bar{b}_s$$

where \bar{b}_s is the mean across category and time of the log biomass in survey s .

As the overall mean of the survey is subtracted, the relative strength of each category within a survey was maintained but the data are now in a standardized format. Plots of the standardized indices by category are presented in Fig. S4.

Analysis

The goal of the analysis was to obtain an overall trend by category over time (Fig. 4E). The data were non-independent at the survey level so a hierarchical approach was adopted that accounts for within-survey correlation.

Linear Mixed Effects with continuous first order within-group correlation

A linear mixed effects model describing the trend for a given category (e.g. Pelagic) was given by:

$$(6) \quad \delta_{s,t}^{Pelagic} = \mu_t + a_s + \varepsilon_{s,t}$$

where μ_t was the yearly fixed effect mean, a_s was the random effect deviation from μ_t by survey s , which were distributed normally $a_s \sim N(0, \sigma_a^2)$, and $\varepsilon_{st} \sim N(0, \sigma_\varepsilon^2)$ were the normally distributed residual errors.

The model described in Equation 6 assumes that the within-group observations are considered exchangeable. To account for the fact that longitudinal data generally have an autocorrelated structure, this basic model was extended to include autocorrelation in the residuals, i.e.

$$(7) \quad \text{Corr}(\varepsilon_{s,t_1}, \varepsilon_{s,t_2}) = \phi^{-|t_2-t_1|}$$

where ϕ was the autocorrelation coefficient.

First-order autocorrelation (AR(1) or a Markov process) occurs when adjacent years are non-independent, with the strength of the dependence decaying with increasing lag time. Note that the correlation structure depends on the time distance between the observations, not on their being strictly consecutive (usual assumption). This maintains the AR(1) correlation structure when missing data values are present. The estimated fixed effects, confidence intervals, and residuals are plotted in Figure S5.

For presentation as a stacked barplot, the standardized trends by category were converted back into positive standardized biomass units using the expected value of a lognormal distribution, including the random effects variance:

$$(8) \quad E[\Delta_t^{Pelagic}] = e^{\mu_t + (\sigma_\varepsilon^2 + \sigma_a^2)/2}$$

where $\Delta_t^{Pelagic}$ is the antilog of $\delta_t^{Pelagic}$. The bias correction for a lognormal variable X when transforming to the original scale is $\text{Var}(X)/2$, here given by $\text{Var}(\Delta_{s,t})/2 = (\sigma_\varepsilon^2 + \sigma_a^2)/2$, assuming the fixed effects values are constant.

Methods to obtain overall trends in $\bar{L}_{\max,t}$

Overall trends in L_{\max} (Fig. 4M-P, Table S6) were based on 19 of the surveys in Table S3 and shown in Table S6. Mean maximum length L_{\max} for each survey-year combination was calculated from:

$$(9) \quad \bar{L}_{\max,s,t} = \frac{\sum_{i=1}^{n_{s,t}} L_{i,s} B_{i,s,t}}{\sum_{i=1}^{n_{s,t}} B_{i,s,t}}$$

where $L_{i,s}$ is the L_{\max} for taxon i in survey region s , $B_{i,s,t}$ is the biomass estimate for taxon i in survey s in year t and $n_{s,t}$ is the number of taxa in survey s in year t . Trends in $\bar{L}_{\max,s,t}$ are presented in Fig. S6 (panel A).

A similar modeling framework to that used for analyzing the standardized biomass indices was implemented to obtain the combined trend of L_{\max} over time. The linear mixed effects model was given by

$$(10) \quad \bar{L}_{\max,s,t} = L_{\max,t} + b_s + \varepsilon_{s,t}$$

where $L_{\max,t}$ is the yearly average, b_s is the deviance from the yearly average by survey s with $b_s \sim N(0, \sigma_b^2)$ and $\varepsilon_{st} \sim N(0, \sigma_\varepsilon^2)$ were the residual errors.

The within-group correlation structure was again assumed AR(1). The average fixed effects, fitted values to each survey and residuals are plotted in Fig. S6 (panels B and C).

The South Georgia Island surveys were not included in the analyses because only five commercially exploited fish species were reported in the source document (S12). Analyses included the Gulf of Thailand surveys only for those years where data were separated out to the level of taxon (18 years, 1963 and 1966-1982). Analyses were based on biomass estimates which were reported for all surveys except for the two University of Rhode Island surveys (Fox Island and Whale Rock), which only reported numbers per survey tow. To avoid over-representing numerous but small taxa for these two surveys, we converted the reported values to a biomass index by multiplying the numbers for all years by the average weight for these species during the later years 1994-2005 when weight data were available. These mean weights were obtained separately for the two University of Rhode Island surveys.

Survey data analyses were conducted in the R statistical programming environment (S13).

Catch data

Mapped global catch-rates (tonnes $\text{km}^{-2} \text{yr}^{-1}$) used in Figs. 1 and 4 and summarized by LME and species group in Fig. S3 were constructed with rule-based procedures developed by the Sea Around Us project (SAUP) based at the Fisheries Centre of the University of British Columbia, Canada (S14) (www.seaaroundus.org, contact Reg Watson) Available fisheries data were harmonized from a wide range of sources including the Food and Agriculture Organization of the UN (FAO) and its regional bodies, the International Council for the Exploration of the Sea (ICES), the Commission for the Conservation of Antarctic Marine Living Resources (CCAMLR), the Northwest Atlantic Fisheries Organization (NAFO), and many reconstructed national datasets (S15) to produce a representative database of global fisheries landings (see also http://www.seaaroundus.org/doc/saup_manual.htm#13). Using additional databases of fishing access arrangements and/or observed national fleet fishing patterns, and extensive information on the distribution and harvest patterns of commercial marine species developed by SAUP, the spatially coarse fisheries landings data records were assigned to a grid of 30-minute latitude \times 30-minute longitude spatial cells from 1950 to 2004. Many subsequent associations were then made possible with this mapped data including global fishing gear use.

Fishing access data

The total years of fishing access by regions (Europe, Russia, Asia) calculated and presented for the 1990s in Fig. 5B were computed by adding the number of years a foreign country had access to EEZs of countries in Africa through a bilateral fishing agreement, as defined by (S16). Where the European Union (EU) entered into an agreement on behalf of its members, the total years were calculated for each member country that gained access rights under the agreement. Details of international fishing agreements were obtained from the Sea Around Us Project (www.seaaroundus.org, contact Dirk Zeller) fishing agreement database, which is based on the FAO FARISIS database corroborated and supplemented using inter-governmental and governmental sources (e.g. EUR-Lex) and other references. Where the terms of an agreement were unknown, duration of one year was used as default. Thus, the information presented here is conservative and excludes illegal access and traditional access that are not formalized through bilateral agreements.

Supporting Tables

Table S1: List of 37 ecosystem models for 31 systems and their sources used to explore multi-species *MSY*. For some systems two EwE models from different time periods were used.

System	Model type	Notes and Source(s)
Alaska Prince William Sound	EwE	(S17)
Aleutians	EwE	(S18)
Australia Darwin Harbour	EwE	(S19)
Baltic	EwE	Database for (S4)
Bay of Biscay	EwE	(S20)
Benguela	2 EwE	Database for (S4)
Black Sea	EwE	Database for (S4)
California Current	2 EwE and 1 Atlantis	(S21, 22)
Canada - Nth Gulf St Lawrence	EwE	(S23)
Central Nth Pacific	EwE	Database for (S4)
Chesapeake	EwE	Database for (S4)
Eastern Bering Sea	EwE	Database for (S4)
Eastern Tropical Pacific	EwE	Database for (S4)
Great Barrier Reef	EwE	(S24)
Georges Bank	EwE	(S25)
Georgia Strait	EwE	Database for (S4)
Gironde Estuary	EwE	(S26)
Grand Banks	EwE	(S27)
Gulf Mexico	EwE	(S28)
Gulf Thailand 1973	EwE	Database for (S4)
Irish Sea	EwE	(S29)
New Zealand	EwE	(S30)
North Sea	EwE	Database for (S4), (S31)
North West Shelf	EwE	(S32)
Port Phillip Bay	EwE	(S33)
SE Alaska 1963	EwE	(S18)
SE Australia	2 EwE and 1 Atlantis	(S34, 35)
Tampa Bay	EwE	Database for (S4)
West Coast Vancouver Island	EwE	Database for (S4)
Western English Channel	2 EwE	(S36)
West Florida Shelf	EwE	(S37)

Table S2. Summary of all stock assessments and their sources used in this analysis and their estimated ratios of current biomass to the equilibrium biomass when harvested at maximum sustainable yield ($B_{current}/B_{MSY}$) and current harvest rate (or fishing mortality rate) to the harvest rate that results in maximum sustainable yield ($u_{current}/u_{MSY}$). The reference ratios were either obtained directly from stock assessments (“Yes”) or from surplus production model fits (“No”); where reference ratios could not be obtained (N/A), the stocks were not plotted in Fig. 3B, but were included in the other analyses.

Large Marine Ecosystem	Scientific name	Fisheries stock	Current year	$B_{current}/B_{MSY}$	$u_{current}/u_{MSY}$	From assessment?	Source
Atlantic High Seas	<i>Thunnus alalunga</i>	Albacore tuna North Atlantic	2005	0.81	1.49	Yes	(S38)
Atlantic High Seas	<i>Thunnus thynnus</i>	Bluefin tuna Eastern Atlantic	2007	0.34	9.38	Yes	(S39)
Baltic Sea	<i>Gadus morhua</i>	Atlantic cod Baltic Areas 22 and 24	2006	0.36	1.43	No	(S40)
Baltic Sea	<i>Gadus morhua</i>	Atlantic cod Baltic Areas 25-32	2006	0.16	1.46	No	(S40)
Baltic Sea	<i>Clupea harengus</i>	Atlantic herring ICES 25-32	2006	0.69	0.79	No	(S40)
Baltic Sea	<i>Clupea harengus</i>	Atlantic herring ICES 30	2006	1.19	1.10	No	(S40)
Baltic Sea	<i>Clupea harengus</i>	Atlantic herring ICES 31	2006	0.29	1.60	No	(S40)
Baltic Sea	<i>Clupea harengus</i>	Atlantic herring ICES 28	2006	1.21	0.87	No	(S40)
Baltic Sea	<i>Sprattus sprattus</i>	Sprat ICES Baltic Areas 22-32	2006	1.13	1.27	No	(S40)
Barents Sea	<i>Gadus morhua</i>	Atlantic cod Northeast Arctic	2006	0.56	1.42	No	(S41)
Barents Sea	<i>Mallotus villosus</i>	Capelin Barents Sea	2006	0.17	0.00	No	(S41)
Barents Sea	<i>Reinhardtius hippoglossoides</i>	Greenland halibut Northeast Arctic	2006	0.36	1.20	No	(S41)
Barents Sea	<i>Melanogrammus aeglefinus</i>	Haddock Northeast Arctic	2006	1.10	1.06	No	(S41)
Barents Sea	<i>Pollachius virens</i>	Saithe Northeast Arctic	2006	1.70	0.60	No	(S41)
Benguela Current	<i>Engraulis encrasicolus</i>	Anchovy South Africa	2006	0.97	0.36	No	(S42)
Benguela Current	<i>Trachurus capensis</i>	Cape horse mackerel South Africa South Coast	2007	1.47	0.76	No	(S43)
Benguela Current	<i>Sardinops sagax</i>	Sardine South Africa	2006	0.75	0.55	No	(S44)
Benguela Current	<i>Palinurus gilchristi</i>	Southern spiny lobster South Africa South Coast	2008	0.51	1.50	No	(S45)
California Current	<i>Reinhardtius stomias</i>	Arrowtooth flounder Pacific Coast	2007	3.81	0.21	Yes	(S46)
California Current	<i>Sebastes melanops</i>	Black rockfish Northern Pacific Coast	2006	1.45	0.53	Yes	(S47)
California Current	<i>Sebastes melanops</i>	Black rockfish Southern Pacific Coast	2007	2.23	0.19	Yes	(S48)
California Current	<i>Sebastes mystinus</i>	Blue rockfish California	2007	0.75	1.55	Yes	(S49)
California Current	<i>Sebastes paucispinis</i>	Bocaccio Southern Pacific Coast	2006	0.32	0.10	Yes	(S50)
California Current	<i>Sebastes pinniger</i>	Canary rockfish Pacific Coast	2007	0.86	0.04	Yes	(S51)
California Current	<i>Sebastes goodei</i>	Chilipepper Southern Pacific Coast	2007	1.96	0.03	Yes	(S52)
California Current	<i>Sebastes levis</i>	Cowcod Southern California	2007	0.09	0.08	Yes	(S53)

California Current	<i>Sebastes crameri</i>	Darkblotched rockfish Pacific Coast	2007	0.73	0.29	Yes	(S54)
California Current	<i>Parophrys vetulus</i>	English sole Pacific Coast	2007	6.42	0.06	Yes	(S55)
California Current	<i>Raja rhina</i>	Longnose skate Pacific Coast	2007	1.76	0.46	Yes	(S56)
California Current	<i>Merluccius productus</i>	Pacific hake Pacific Coast	2008	1.61	0.73	Yes	(S57)
California Current	<i>Sebastes alutus</i>	Pacific ocean perch Pacific Coast	2007	0.69	0.08	Yes	(S58)
California Current	<i>Anoplopoma fimbria</i>	Sablefish Pacific Coast	2007	1.02	0.69	Yes	(S59)
California Current	<i>Sebastes entomelas</i>	Widow rockfish Pacific Coast	2006	0.88	0.05	Yes	(S60)
California Current	<i>Sebastes ruberrimus</i>	Yelloweye rockfish Pacific Coast	2006	0.83	0.61	Yes	(S61)
Celtic-Biscay Shelf	<i>Gadus morhua</i>	Atlantic cod Irish Sea	2006	0.15	0.56	No	(S62)
Celtic-Biscay Shelf	<i>Gadus morhua</i>	Atlantic cod West of Scotland	2006	0.12	0.42	No	(S62)
Celtic-Biscay Shelf	<i>Micromesistius poutassou</i>	Blue whiting Northeast Atlantic	2006	0.67	1.66	No	(S63)
Celtic-Biscay Shelf	<i>Solea vulgaris</i>	Common European sole Bay of Biscay	2006	0.75	1.00	No	(S64)
Celtic-Biscay Shelf	<i>Solea vulgaris</i>	Common European sole Irish Sea	2006	0.36	1.16	No	(S62)
Celtic-Biscay Shelf	<i>Solea vulgaris</i>	Common European sole ICES VIIId	2006	1.41	0.68	No	(S65)
Celtic-Biscay Shelf	<i>Solea vulgaris</i>	Common European sole Celtic Sea	2006	0.90	0.95	No	(S66)
Celtic-Biscay Shelf	<i>Solea vulgaris</i>	Common European sole Western English Channel	2006	0.51	1.74	No	(S66)
Celtic-Biscay Shelf	<i>Pleuronectes platessa</i>	European plaice Irish Sea	2006	1.07	0.23	No	(S62)
Celtic-Biscay Shelf	<i>Pleuronectes platessa</i>	European plaice ICES VIIe-k	2006	0.65	0.41	No	(S66)
Celtic-Biscay Shelf	<i>Pleuronectes platessa</i>	European plaice ICES VIIe	2006	0.51	1.39	No	(S66)
Celtic-Biscay Shelf	<i>Melanogrammus aeglefinus</i>	Haddock West of Scotland	2006	0.58	0.73	No	(S62)
Celtic-Biscay Shelf	<i>Merluccius merluccius</i>	Hake Northeast Atlantic North	2006	1.04	0.74	No	(S64)
Celtic-Biscay Shelf	<i>Clupea harengus</i>	Atlantic herring Northern Irish Sea	2006	0.72	0.34	No	(S67)
Celtic-Biscay Shelf	<i>Clupea harengus</i>	Atlantic herring ICES VIa	2006	0.18	1.59	No	(S67)
Celtic-Biscay Shelf	<i>Clupea harengus</i>	Atlantic herring ICES VIa-VIIb-VIIc	2000	0.50	1.04	No	(S67)
Celtic-Biscay Shelf	<i>Scomber scombrus</i>	Mackerel ICES Northeast Atlantic	2006	0.98	0.73	No	(S68)
Celtic-Biscay Shelf	<i>Merlangius merlangus</i>	Whiting ICES VIIe-k	2006	0.44	1.25	No	(S66)
Eastern Bering Sea	<i>Pleuronectes quadrituberculatus</i>	Alaska plaice Bering Sea and Aleutian Islands	2007	2.20	0.06	Yes	(S69)
Eastern Bering Sea	<i>Reinhardtius stomias</i>	Arrowtooth flounder Bering Sea and Aleutian Islands	2008	2.70	0.31	No	(S70)
Eastern Bering Sea	<i>Pleurogrammus monopterygius</i>	Atka mackerel Bering Sea and Aleutian Islands	2008	1.71	0.55	No	(S71)
Eastern Bering Sea	<i>Hippoglossoides elassodon</i>	Flathead sole Bering Sea and Aleutian Islands	2008	1.83	0.18	No	(S72)
Eastern Bering Sea	<i>Reinhardtius hippoglossoides</i>	Greenland turbot Bering Sea and Aleutian Islands	2007	1.46	0.05	Yes	(S73)
Eastern Bering Sea	<i>Lepidopsetta polyxystra</i>	Northern rock sole Eastern Bering Sea and Aleutian Islands	2007	3.02	0.21	Yes	(S74)
Eastern Bering Sea	<i>Sebastes polyspinis</i>	Northern rockfish Bering Sea and Aleutian Islands	2008	1.42	0.13	No	(S75)

Eastern Bering Sea	<i>Gadus macrocephalus</i>	Pacific cod Bering Sea and Aleutian Islands	2007	1.14	0.93	No	(S76)
Eastern Bering Sea	<i>Sebastes alutus</i>	Pacific ocean perch Eastern Bering Sea and Aleutian Islands	2008	1.27	0.26	No	(S77)
Eastern Bering Sea	<i>Chionoecetes opilio</i>	Snow crab Bering Sea	2008	0.55	1.49	No	(S78)
Eastern Bering Sea	<i>Chionoecetes bairdi</i>	Tanner crab Bering Sea	2007	0.79	0.15	No	(S78)
Eastern Bering Sea	<i>Theragra chalcogramma</i>	Walleye pollock Eastern Bering Sea	2007	0.92	0.94	No	(S79)
Eastern Bering Sea	<i>Limanda aspera</i>	Yellowfin sole Bering Sea and Aleutian Islands	2007	2.00	0.69	Yes	(S80)
Eastern Bering Sea	<i>Paralithodes platypus</i>	Blue king crab Pribilof Islands	2008	0.08	0.00	Yes	(S78)
Eastern Bering Sea	<i>Paralithodes camtschaticus</i>	Red king crab Norton Sound	2008	1.47	NA	Yes	(S78)
Eastern Bering Sea	<i>Paralithodes platypus</i>	Blue king crab St. Matthew Island	2008	1.45	NA	Yes	(S78)
Eastern Bering Sea	<i>Paralithodes camtschaticus</i>	Red king crab Pribilof Islands	2009	1.44	NA	Yes	(S78)
Eastern Bering Sea	<i>Paralithodes camtschaticus</i>	Red king crab Bristol Bay	2008	1.27	1.05	Yes	(S78)
Eastern Bering Sea	<i>Lithodes aequispinus</i>	Golden king crab Aleutian Islands Eastern Stock	2007	0.61	NA	Yes	(S78)
Eastern Bering Sea	<i>Lithodes aequispinus</i>	Golden king crab Aleutian Islands Western Stock	2007	0.53	NA	Yes	(S78)
Faroe Plateau	<i>Gadus morhua</i>	Atlantic cod Faroe Plateau	2006	0.26	1.52	No	(S81)
Faroe Plateau	<i>Melanogrammus aeglefinus</i>	Haddock Faroe Plateau	2006	0.85	1.07	No	(S81)
Faroe Plateau	<i>Pollachius virens</i>	Saithe Faroe Plateau	2006	0.99	1.52	No	(S81)
Gulf of Alaska	<i>Parophrys vetulus</i>	English sole Hecate Strait	2001	1.23	0.37	No	(S82)
Gulf of Alaska	<i>Gadus macrocephalus</i>	Pacific cod Hecate Strait	2004	1.08	0.18	No	(S83)
Gulf of Alaska	<i>Gadus macrocephalus</i>	Pacific cod West Coast of Vancouver Island	2001	1.04	0.47	Yes	(S84)
Gulf of Alaska	<i>Clupea pallasii</i>	Pacific herring Central Coast	2007	0.30	0.11	No	(S85)
Gulf of Alaska	<i>Clupea pallasii</i>	Pacific herring Prince Rupert District	2007	0.16	0.32	No	(S85)
Gulf of Alaska	<i>Clupea pallasii</i>	Pacific herring Queen Charlotte Islands	2007	0.20	0.00	No	(S85)
Gulf of Alaska	<i>Clupea pallasii</i>	Pacific herring Strait of Georgia	2007	0.91	0.40	No	(S85)
Gulf of Alaska	<i>Clupea pallasii</i>	Pacific herring West Coast of Vancouver Island	2007	0.03	0.00	No	(S85)
Gulf of Alaska	<i>Lepidopsetta bilineata</i>	Rock sole Hecate Strait	2001	1.03	0.45	No	(S82)
Gulf of Alaska	<i>Anoplopoma fimbria</i>	Sablefish Eastern Bering Sea / Aleutian Islands / Gulf of Alaska	2007	1.05	0.66	Yes	(S86)
Gulf of Mexico	<i>Mycteroperca microlepis</i>	Gag Gulf of Mexico	2004	1.00	1.99	Yes	(S87)
Gulf of Mexico	<i>Brevoortia patronus</i>	Gulf menhaden Gulf of Mexico	2004	1.08	0.48	No	(S88)
Iberian Coastal	<i>Lepidorhombus boscii</i>	Fourspotted megrim ICES VIIIc-IXa	2006	0.70	1.01	No	(S64)
Iberian Coastal	<i>Lepidorhombus whiffiagonis</i>	Megrim ICES VIIIc-IXa	2006	0.43	1.07	No	(S64)
Iceland Shelf	<i>Gadus morhua</i>	Atlantic cod Iceland	2006	0.46	1.17	No	(S81)
Iceland Shelf	<i>Mallotus villosus</i>	Capelin Iceland	2006	0.49	0.85	No	(S81)

Iceland Shelf	<i>Melanogrammus aeglefinus</i>	Haddock Iceland	2007	0.98	1.23	No	(S81)
Iceland Shelf	<i>Clupea harengus</i>	Atlantic herring Iceland (summer spawners)	2006	1.00	0.79	No	(S81)
Mediterranean Sea	<i>Xiphias gladius</i>	Swordfish Mediterranean Sea	2005	0.94	1.26	Yes	(S89)
New Zealand Shelf	<i>Genypterus blacodes</i>	Ling New Zealand Areas LIN3 and LIN4	2007	3.07	0.09	Yes	(S90)
New Zealand Shelf	<i>Genypterus blacodes</i>	Ling New Zealand Areas LIN5 and LIN6	2007	3.96	0.10	Yes	(S90)
New Zealand Shelf	<i>Genypterus blacodes</i>	Ling New Zealand Area LIN6b	2006	2.19	0.11	Yes	(S91)
New Zealand Shelf	<i>Genypterus blacodes</i>	Ling New Zealand Area LIN7CK	2007	2.49	0.32	Yes	(S90)
New Zealand Shelf	<i>Genypterus blacodes</i>	Ling New Zealand Area LIN7WC	2008	2.21	0.13	Yes	(S91)
New Zealand Shelf	<i>Allocyttus niger</i>	Black oreo west end of Chatham Rise	2007	0.99	0.82	Yes	(S91)
New Zealand Shelf	<i>Haliotis iris</i>	Paua New Zealand Area PAU5A	2006	0.72	2.83	No	(S92)
New Zealand Shelf	<i>Haliotis iris</i>	Paua New Zealand Area PAU5B	2007	1.02	0.59	No	(S93)
New Zealand Shelf	<i>Haliotis iris</i>	Paua New Zealand Area PAU5D	2006	0.44	2.10	No	(S92)
New Zealand Shelf	<i>Haliotis iris</i>	Paua New Zealand Area PAU7	2008	0.87	0.94	No	(S94)
New Zealand Shelf	<i>Rexea solandri</i>	Common gemfish	2006	1.61	0.30	Yes	(S95)
New Zealand Shelf	<i>Macruronus novaezelandiae</i>	Hoki Eastern New Zealand	2007	1.11	0.33	No	(S96)
New Zealand Shelf	<i>Macruronus novaezelandiae</i>	Hoki Western New Zealand	2007	0.51	0.57	No	(S96)
New Zealand Shelf	<i>Chrysophrys auratus</i>	New Zealand snapper New Zealand SNA8	2005	0.35	2.50	Yes	(S97)
New Zealand Shelf	<i>Pseudocyttus maculatus</i>	Smooth oreo west end of Chatham Rise	2004	1.06	0.54	No	(S91)
New Zealand Shelf	<i>Micromesistius australis</i>	Southern blue whiting Campbell Island Rise	2006	0.86	1.20	No	(S98)
New Zealand Shelf	<i>Merluccius australis</i>	Southern hake Chatham Rise	2006	1.77	0.12	Yes	(S99)
New Zealand Shelf	<i>Merluccius australis</i>	Southern hake Sub-Antarctic	2007	2.91	0.11	Yes	(S100)
New Zealand Shelf	<i>Pseudocaranx dentex</i>	Trevally New Zealand Area TRE7	2005	1.44	0.83	Yes	(S101)
Newfoundland-Labrador Shelf	<i>Hippoglossoides platessoides</i>	American plaice NAFO 23K	2003	0.12	0.07	No	(S102)
Newfoundland-Labrador Shelf	<i>Hippoglossoides platessoides</i>	American plaice NAFO 3LNO	2006	0.08	0.77	No	(S103)
Newfoundland-Labrador Shelf	<i>Gadus morhua</i>	Atlantic cod NAFO 3Ps	2004	0.48	0.41	No	(S104)
Newfoundland-Labrador Shelf	<i>Gadus morhua</i>	Atlantic cod NAFO 3Pn4RS	2006	0.09	0.79	No	(S105)
Newfoundland-Labrador Shelf	<i>Gadus morhua</i>	Atlantic cod NAFO 3NO	2006	0.02	0.27	No	(S106)
Newfoundland-Labrador Shelf	<i>Reinhardtius hippoglossoides</i>	Greenland halibut NAFO 23KLMNO	2006	0.39	1.73	No	(S107)
Newfoundland-Labrador Shelf	<i>Redfish species</i>	Redfish species NAFO 3LN	2006	1.91	0.01	Yes	(S108)
North Sea	<i>Gadus morhua</i>	Atlantic cod Kattegat	2006	0.19	0.31	No	(S40)
North Sea	<i>Gadus morhua</i>	Atlantic cod North Sea	2006	0.19	0.80	No	(S65)
North Sea	<i>Solea vulgaris</i>	Common European sole ICES Kattegat and Skagerrak	2006	1.25	0.54	No	(S40)
North Sea	<i>Melanogrammus aeglefinus</i>	Haddock ICES IIIa and North Sea	2006	0.62	0.25	No	(S65)

North Sea	<i>Clupea harengus</i>	Atlantic herring North Sea	2006	0.65	1.32	No	(\$67)
North Sea	<i>Trisopterus esmarkii</i>	Norway pout North Sea	2006	0.90	0.33	No	(\$65)
North Sea	<i>Pollachius virens</i>	Saithe ICES IIIa, VI and North Sea	2006	0.57	0.97	No	(\$65)
North Sea	<i>Ammodytes marinus</i>	Sandeel North Sea	2007	0.92	0.24	No	(\$65)
North Sea	<i>Merlangius merlangus</i>	Whiting ICES IIIa, VIIId and North Sea	2006	0.33	1.04	No	(\$65)
Northeast U.S. Shelf	<i>Homarus americanus</i>	American lobster Rhode Island	2006	0.61	0.73	Yes	(\$109)
Northeast U.S. Shelf	<i>Hippoglossoides platessoides</i>	American plaice NAFO 5YZ	2007	0.70	0.30	No	(\$110)
Northeast U.S. Shelf	<i>Gadus morhua</i>	Atlantic cod NAFO 5Zjm	2002	0.34	0.45	No	(\$111)
Northeast U.S. Shelf	<i>Gadus morhua</i>	Atlantic cod Georges Bank	2007	0.12	0.72	No	(\$110)
Northeast U.S. Shelf	<i>Gadus morhua</i>	Atlantic cod Gulf of Maine	2007	0.63	2.40	Yes	(\$110)
Northeast U.S. Shelf	<i>Melanogrammus aeglefinus</i>	Haddock NAFO 4X5Y	2003	0.85	0.33	No	(\$112)
Northeast U.S. Shelf	<i>Melanogrammus aeglefinus</i>	Haddock NAFO 5Zejm	2002	1.00	0.65	No	(\$113)
Northeast U.S. Shelf	<i>Melanogrammus aeglefinus</i>	Haddock NAFO 5Y	2007	0.99	1.21	No	(\$110)
Northeast U.S. Shelf	<i>Pollachius virens</i>	Pollock NAFO 4VWX5Zc	2006	0.56	0.30	No	(\$114)
Northeast U.S. Shelf	<i>Tautoga onitis</i>	Tautog Rhode Island	2006	0.79	0.62	Yes	(\$109)
Northeast U.S. Shelf	<i>Pseudopleuronectes americanus</i>	Winter flounder Southern New England-Mid Atlantic	2007	0.09	1.10	No	(\$110)
Northeast U.S. Shelf	<i>Pseudopleuronectes americanus</i>	Winter flounder Rhode Island	2006	0.23	2.02	Yes	(\$109)
Northeast U.S. Shelf	<i>Limanda ferruginea</i>	Yellowtail flounder Georges Bank	2007	0.22	1.14	Yes	(\$110)
Northeast U.S. Shelf	<i>Brevoortia tyrannus</i>	Atlantic menhaden	2005	0.47	0.97	No	(\$115)
Norwegian Sea	<i>Gadus morhua</i>	Atlantic cod coastal Norway	2006	0.27	2.17	No	(\$41)
Pacific High Seas	<i>Thunnus alalunga</i>	Albacore tuna South Pacific Ocean	2006	2.46	0.91	Yes	(\$116)
Pacific High Seas	<i>Thunnus obesus</i>	Bigeye tuna Western Pacific Ocean	2006	1.05	1.38	Yes	(\$117)
Pacific High Seas	<i>Katsuwonus pelamis</i>	Skipjack tuna Central Western Pacific	2006	4.38	0.31	Yes	(\$118)
Pacific High Seas	<i>Thunnus albacares</i>	Yellowfin tuna Central Western Pacific	2005	1.22	0.80	Yes	(\$119)
Patagonian Shelf	<i>Merluccius hubbsi</i>	Argentine hake Northern Argentina	2007	0.19	1.26	Yes	(\$120)
Patagonian Shelf	<i>Merluccius hubbsi</i>	Argentine hake Southern Argentina	2007	0.54	1.67	Yes	(\$121)
Patagonian Shelf	<i>Macruronus magellanicus</i>	Patagonian grenadier Southern Argentina	2006	2.15	0.60	Yes	(\$122)
Patagonian Shelf	<i>Micromesistius australis</i>	Southern blue whiting Southern Argentina	2007	0.38	1.18	No	(\$123)
Scotian Shelf	<i>Gadus morhua</i>	Atlantic cod NAFO 4TVn	2006	0.17	0.32	No	(\$124)
Southern Australian Shelf	<i>Genypterus blacodes</i>	Ling Great Australian Bight	2007	1.08	8.98	No	(\$125)
Southern Australian Shelf	<i>Genypterus blacodes</i>	Ling Southeast Australia	2007	0.59	2.20	No	(\$125)
Southern Australian Shelf	<i>Seriolella brama</i>	Blue warehou Great Australian Bight	2006	0.41	2.04	No	(\$126)
Southern Australian Shelf	<i>Seriolella brama</i>	Blue warehou Southeast Australia	2006	0.49	0.84	No	(\$126)

Southern Australian Shelf	<i>Rexea solandri</i>	Common gemfish Southeast Australia	2007	0.25	0.39	No	(S127)
Southern Australian Shelf	<i>Platycephalus conatus</i>	Deepwater flathead Southeast Australia	2006	1.43	0.61	No	(S128)
Southern Australian Shelf	<i>Nemadactylus macropterus</i>	Jackass morwong Southeast Australia	2007	0.31	1.80	No	(S129)
Southern Australian Shelf	<i>Hoplostethus atlanticus</i>	Orange roughy Southeast Australia	2006	0.48	0.29	No	(S130)
Southern Australian Shelf	<i>Sillago flindersi</i>	School whiting Southeast Australia	2007	0.66	0.82	No	(S131)
Southern Australian Shelf	<i>Seriolella punctata</i>	Silverfish Southeast Australia	2006	1.03	0.79	No	(S132)
Southern Australian Shelf	<i>Neoplatycephalus richardsoni</i>	Tiger flathead Southeast Australia	2006	1.78	1.03	No	(S133)
Southeast U.S. Shelf	<i>Pagrus pagrus</i>	Red porgy Southern Atlantic coast	2004	0.61	0.39	Yes	(S134)
Southeast U.S. Shelf	<i>Scomberomorus maculatus</i>	Spanish mackerel Southern Atlantic Coast	2007	0.47	0.91	Yes	(S135)

Table S3. Summary of the trawl surveys compiled for analysis, the number of years in which surveys were conducted, the time span of the surveys, the number of taxa included in the analysis, how many of these taxa were identified to species, and the number of taxa that were invertebrate, pelagic (mid-water fish species) and demersal (bottom-dwelling fish species). NR = taxa not reported in a given survey, URI = University of Rhode Island, MLI = Maurice Lamontagne Institute, NOAA = National Oceanic and Atmospheric Administration, CEFAS = Centre for Environment, Fisheries and Aquaculture Science, CSIRO = Commonwealth Scientific and Industrial Research Organisation, ADF&G = Alaska Department of Fish and Game.

Survey name	Years	Year range	Taxa	Species	Invertebrates	Pelagics	Demersals	Source or analyst
St. Pierre Bank, Newfoundland	40	1951-1995	27	26	NR	1	26	(S136)
Southern Grand Banks, Newfoundland	41	1952-1995	20	19	NR	1	19	(S136)
Northern Gulf of St. Lawrence	18	1990-2007	10	9	NR	1	9	Diane Archambault, MLI
Southern Gulf of St. Lawrence	37	1971-2007	52	49	NR	7	45	(S137)
Scotian Shelf and Bay of Fundy	36	1970-2006	49	48	1	6	42	(S138)
Gulf of Maine	45	1963-2007	34	34	3	4	27	Michael Fogarty, NOAA
Georges Bank	45	1963-2007	40	40	3	1	36	Michael Fogarty, NOAA
URI Fox Island	47	1959-2005	25	23	7	2	16	(S139)
URI Whale Rock	47	1959-2005	25	23	7	2	16	(S139)
Mid-Atlantic Bight	41	1967-2007	40	40	3	4	33	Michael Fogarty, NOAA
South Georgia, Subantarctic	23	1970-1992	5	5	NR	1	4	(S12)
Celtic Sea	18	1987-2004	55	53	NR	9	46	Simon Jennings, CEFAS
North Sea	28	1980-2007	49	53	NR	4	45	Simon Jennings, CEFAS
North-west Australia	13	1978-1997	542	538	5	39	498	Beth Fulton, CSIRO
Gulf of Thailand	35	1961-1995	38	5	6	6	26	(S140, S141)
Eastern Bering Sea	27	1982-2008	32	12	13	2	17	Robert Lauth, NOAA
Aleutian Islands	10	1980-2006	66	58	2	2	62	Mark Wilkins, NOAA
Gulf of Alaska small mesh	36	1972-2007	24	13	11	3	10	Aaren Ellsworth, ADF&G
Gulf of Alaska trawl	10	1984-2007	118	105	6	8	104	Mark Wilkins, NOAA
US West Coast	10	1977-2004	58	57	1	9	48	Mark Wilkins, NOAA

Table S4. Ecosystem exploitation rates ($u_{ave}=C_{tot}/B_{tot}$) and average ratio of biomass to B_{MSY} ($B_{ave} = B/B_{MSY}$) for ecosystems plotted in Figure 3A, based on stock assessments. The average B/B_{MSY} ratio is the geometric mean of the ratios for individual fishery stocks. For this analysis, Pacific hake, Atlantic menhaden, and blue whiting were excluded from the California Current, Northeast U.S. Shelf, and Celtic-Biscay Shelf, respectively (see above text for explanation).

Year	Iceland Shelf		North Sea		Celtic-Biscay Shelf		S. Australia Shelf		California Current		Northeast U.S. Shelf		Newfoundland-Labrad.		Baltic Sea		Eastern Bering Sea		New Zealand Shelf	
	u_{ave}	B_{ave}	u_{ave}	B_{ave}	u_{ave}	B_{ave}	u_{ave}	B_{ave}	u_{ave}	B_{ave}	u_{ave}	B_{ave}	u_{ave}	B_{ave}	u_{ave}	B_{ave}	u_{ave}	B_{ave}	u_{ave}	B_{ave}
1975									0.033	1.853			0.174	0.538	0.164	0.803			0.017	2.728
1976									0.050	1.810			0.176	0.532	0.144	0.857			0.024	2.641
1977									0.030	1.760			0.169	0.574	0.154	0.744			0.038	2.479
1978					0.186	0.649	0.030	2.030	0.038	1.746	0.325	0.647	0.128	0.861	0.144	0.700	0.140	0.792	0.009	2.451
1979	0.380	0.771			0.218	0.787	0.027	2.029	0.057	1.707	0.283	0.672	0.151	0.800	0.156	0.676	0.121	0.813	0.015	2.246
1980	0.415	0.673			0.211	0.841	0.032	2.013	0.062	1.623	0.345	0.783	0.149	0.829	0.186	0.672	0.105	0.922	0.013	2.284
1981	0.464	0.633			0.216	0.871	0.030	2.029	0.074	1.543	0.329	0.678	0.150	0.877	0.171	0.693	0.079	1.020	0.014	2.185
1982	0.210	0.555			0.224	0.785	0.029	1.964	0.090	1.426	0.379	0.690	0.144	0.904	0.156	0.703	0.070	1.067	0.014	2.098
1983	0.311	0.614	0.221	0.917	0.192	0.836	0.029	1.900	0.067	1.278	0.382	0.628	0.133	0.954	0.146	0.817	0.068	0.977	0.016	2.046
1984	0.355	0.623	0.259	0.804	0.201	0.863	0.027	1.902	0.061	1.169	0.346	0.540	0.131	0.941	0.180	0.773	0.079	0.868	0.019	1.973
1985	0.439	0.634	0.285	0.697	0.181	0.856	0.030	1.777	0.062	1.069	0.377	0.448	0.142	0.931	0.187	0.729	0.077	0.744	0.015	1.981
1986	0.403	0.571	0.229	0.874	0.183	0.837	0.035	2.002	0.061	0.963	0.330	0.417	0.186	0.852	0.197	0.624	0.080	0.730	0.026	1.966
1987	0.378	0.611	0.247	0.742	0.198	0.918	0.040	1.981	0.068	0.893	0.343	0.361	0.220	0.716	0.180	0.668	0.063	0.854	0.045	1.942
1988	0.411	0.635	0.338	0.583	0.203	0.894	0.050	2.010	0.061	0.852	0.362	0.338	0.189	0.624	0.195	0.625	0.091	0.920	0.078	1.907
1989	0.424	0.581	0.297	0.632	0.181	0.772	0.132	1.918	0.071	0.810	0.299	0.344	0.204	0.550	0.179	0.610	0.091	1.007	0.048	2.007
1990	0.304	0.503	0.251	0.617	0.208	0.704	0.175	1.771	0.069	0.772	0.371	0.351	0.229	0.461	0.151	0.652	0.111	1.183	0.048	1.989
1991	0.356	0.512	0.274	0.670	0.204	0.660	0.143	1.535	0.065	0.739	0.429	0.340	0.289	0.358	0.132	0.627	0.119	1.255	0.063	1.920
1992	0.375	0.491	0.291	0.725	0.222	0.679	0.166	1.304	0.061	0.712	0.430	0.297	0.293	0.243	0.116	0.649	0.108	1.361	0.082	1.796
1993	0.444	0.551	0.292	0.646	0.246	0.657	0.113	1.129	0.059	0.678	0.419	0.255	0.311	0.175	0.123	0.701	0.089	1.351	0.061	1.856
1994	0.348	0.545	0.256	0.761	0.252	0.649	0.090	1.018	0.051	0.654	0.361	0.233	0.188	0.128	0.165	0.733	0.091	1.394	0.052	1.860
1995	0.309	0.547	0.251	0.786	0.230	0.636	0.088	0.952	0.049	0.644	0.249	0.257	0.055	0.127	0.143	0.732	0.081	1.379	0.055	1.786
1996	0.475	0.514	0.251	0.645	0.188	0.619	0.085	0.932	0.051	0.643	0.252	0.290	0.059	0.150	0.184	0.706	0.086	1.401	0.060	1.758
1997	0.433	0.489	0.194	0.726	0.181	0.618	0.089	0.987	0.055	0.643	0.261	0.314	0.087	0.163	0.221	0.611	0.094	1.242	0.088	1.634
1998	0.434	0.484	0.281	0.553	0.214	0.593	0.087	0.932	0.040	0.644	0.253	0.321	0.100	0.176	0.219	0.559	0.089	1.178	0.106	1.561
1999	0.379	0.485	0.184	0.633	0.205	0.531	0.086	0.795	0.046	0.659	0.219	0.322	0.125	0.188	0.234	0.526	0.076	1.100	0.099	1.525
2000	0.447	0.457	0.167	0.667	0.233	0.544	0.091	0.685	0.039	0.678	0.211	0.357	0.137	0.188	0.214	0.570	0.084	0.833	0.111	1.454
2001	0.448	0.511	0.239	0.529	0.241	0.586	0.083	0.636	0.032	0.708	0.212	0.397	0.124	0.192	0.223	0.525	0.097	1.131	0.141	1.366
2002	0.382	0.618	0.198	0.540	0.279	0.580	0.094	0.632	0.021	0.764	0.193	0.374	0.118	0.172	0.214	0.510	0.100	1.134	0.106	1.381

Table S5. Annual total biomass (metric tons) from stock assessments for all data and for the three focal regions, corresponding to Figs 4A-D. This is based on stocks with assessment biomass data for at least 25 years within 1977-2006.

Region	Year	Demersal ≥90	Demersal 30-90	Demersal ≤30	Invertebrate	Pelagic	<i>n</i> stocks
All data	1977	17,816,886	10,933,201	-	886,431	39,198,859	144
All data	1978	18,148,804	11,277,329	-	819,442	40,085,956	144
All data	1979	18,621,668	11,417,867	-	795,435	38,718,873	144
All data	1980	20,975,601	12,076,138	-	772,254	39,829,121	144
All data	1981	24,117,281	13,029,132	-	697,201	36,062,916	144
All data	1982	24,681,175	13,750,325	-	554,708	35,554,985	144
All data	1983	26,628,867	14,669,171	-	536,099	44,591,334	144
All data	1984	25,715,680	14,709,692	-	544,492	42,056,401	144
All data	1985	27,878,462	15,070,862	-	557,139	40,471,702	144
All data	1986	28,373,242	14,926,243	-	705,275	36,382,584	144
All data	1987	26,844,578	15,595,944	-	865,234	36,826,942	144
All data	1988	23,972,032	15,064,374	-	1,015,574	33,677,846	144
All data	1989	23,091,704	15,188,085	-	1,123,676	32,101,630	144
All data	1990	21,028,206	14,909,796	-	1,141,282	36,447,263	144
All data	1991	18,285,742	14,697,201	-	990,974	38,417,689	144
All data	1992	20,657,921	14,625,057	-	841,230	36,519,227	144
All data	1993	24,018,728	14,550,552	-	721,690	32,644,692	144
All data	1994	24,750,393	13,897,749	-	691,168	30,501,946	144
All data	1995	26,049,168	13,372,661	-	696,255	30,510,852	144
All data	1996	24,255,480	12,882,579	-	692,830	29,314,150	144
All data	1997	21,573,705	12,842,952	-	601,987	31,217,521	144
All data	1998	20,356,689	12,584,281	-	431,302	32,884,159	144
All data	1999	23,209,113	12,580,708	-	307,177	35,375,669	144
All data	2000	22,251,220	13,317,021	-	276,214	35,606,766	144
All data	2001	20,031,195	13,610,095	-	263,017	35,958,787	144
All data	2002	21,032,807	14,482,568	-	265,593	36,687,404	144
All data	2003	21,976,779	14,545,071	-	283,830	34,845,474	144
All data	2004	21,470,974	14,433,304	-	312,984	33,391,206	144
All data	2005	21,301,002	14,068,621	-	351,668	30,066,567	144
All data	2006	17,885,573	13,890,451	-	410,944	29,066,058	144
Eastern Bering Sea	1977	4,182,241	4,828,054	-	875,574	-	15
Eastern Bering Sea	1978	4,101,016	5,216,561	-	809,077	-	15
Eastern Bering Sea	1979	4,142,030	5,538,150	-	785,558	-	15
Eastern Bering Sea	1980	5,526,030	6,001,714	-	762,856	-	15
Eastern Bering Sea	1981	9,636,570	6,271,765	-	687,900	-	15
Eastern Bering Sea	1982	11,059,670	6,474,280	-	545,966	-	15
Eastern Bering Sea	1983	12,222,520	6,673,643	-	527,850	-	15
Eastern Bering Sea	1984	11,921,580	6,868,758	-	536,373	-	15
Eastern Bering Sea	1985	14,165,850	6,897,416	-	549,137	-	15
Eastern Bering Sea	1986	13,368,060	6,893,876	-	696,990	-	15
Eastern Bering Sea	1987	13,994,680	7,114,904	-	856,804	-	15
Eastern Bering Sea	1988	13,192,110	7,187,231	-	1,006,658	-	15
Eastern Bering Sea	1989	11,312,030	7,430,937	-	1,113,805	-	15
Eastern Bering Sea	1990	9,107,410	7,763,384	-	1,130,892	-	15

Eastern Bering Sea	1991	7,153,070	7,988,098	-	980,816	-	15
Eastern Bering Sea	1992	10,416,780	8,428,810	-	831,412	-	15
Eastern Bering Sea	1993	12,634,860	8,468,827	-	711,773	-	15
Eastern Bering Sea	1994	12,310,360	8,410,467	-	681,495	-	15
Eastern Bering Sea	1995	14,293,790	8,316,493	-	686,104	-	15
Eastern Bering Sea	1996	12,398,500	8,207,331	-	682,239	-	15
Eastern Bering Sea	1997	10,918,220	8,053,359	-	592,024	-	15
Eastern Bering Sea	1998	10,916,180	7,768,184	-	421,909	-	15
Eastern Bering Sea	1999	11,901,510	7,644,098	-	298,740	-	15
Eastern Bering Sea	2000	11,098,780	7,523,617	-	268,622	-	15
Eastern Bering Sea	2001	10,816,550	7,501,460	-	256,057	-	15
Eastern Bering Sea	2002	11,256,650	7,549,053	-	259,051	-	15
Eastern Bering Sea	2003	12,889,180	7,716,193	-	277,267	-	15
Eastern Bering Sea	2004	11,904,540	8,005,913	-	306,161	-	15
Eastern Bering Sea	2005	9,914,060	8,093,434	-	344,444	-	15
Eastern Bering Sea	2006	7,671,050	8,192,378	-	403,452	-	15
Eastern Bering Sea	2007	6,180,190	8,348,116	-	443,414	-	15
Eastern Bering Sea	2008	5,297,950	8,303,982	-	426,422	-	15
Eastern Canada	1977	716,384	1,033,571	-	-	-	8
Eastern Canada	1978	918,804	1,043,653	-	-	-	8
Eastern Canada	1979	939,788	905,712	-	-	-	8
Eastern Canada	1980	1,046,927	916,708	-	-	-	8
Eastern Canada	1981	1,144,446	906,104	-	-	-	8
Eastern Canada	1982	1,171,863	939,001	-	-	-	8
Eastern Canada	1983	1,276,854	961,264	-	-	-	8
Eastern Canada	1984	1,272,302	888,437	-	-	-	8
Eastern Canada	1985	1,329,463	853,920	-	-	-	8
Eastern Canada	1986	1,301,265	774,311	-	-	-	8
Eastern Canada	1987	957,556	701,209	-	-	-	8
Eastern Canada	1988	791,766	645,418	-	-	-	8
Eastern Canada	1989	639,677	604,011	-	-	-	8
Eastern Canada	1990	556,974	513,773	-	-	-	8
Eastern Canada	1991	402,520	472,523	-	-	-	8
Eastern Canada	1992	274,751	378,716	-	-	-	8
Eastern Canada	1993	183,943	299,782	-	-	-	8
Eastern Canada	1994	166,161	229,391	-	-	-	8
Eastern Canada	1995	191,785	215,451	-	-	-	8
Eastern Canada	1996	227,501	242,546	-	-	-	8
Eastern Canada	1997	256,522	264,043	-	-	-	8
Eastern Canada	1998	261,392	306,123	-	-	-	8
Eastern Canada	1999	269,643	341,165	-	-	-	8
Eastern Canada	2000	247,759	366,462	-	-	-	8
Eastern Canada	2001	248,661	377,957	-	-	-	8
Eastern Canada	2002	233,626	368,800	-	-	-	8
Eastern Canada	2003	218,810	377,914	-	-	-	8
Eastern Canada	2004	233,618	373,657	-	-	-	8
Eastern Canada	2005	235,779	382,918	-	-	-	8
Eastern Canada	2006	226,761	387,477	-	-	-	8
Northeast U.S. Shelf	1978	399,970	104,135	-	1,333	-	14
Northeast U.S. Shelf	1979	414,797	106,430	-	1,077	-	14

Northeast U.S. Shelf	1980	435,759	108,736	-	992	-	14
Northeast U.S. Shelf	1981	428,435	102,113	-	1,530	-	14
Northeast U.S. Shelf	1982	415,449	99,468	-	1,555	-	14
Northeast U.S. Shelf	1983	367,021	92,828	-	1,531	-	14
Northeast U.S. Shelf	1984	324,101	74,021	-	1,780	-	14
Northeast U.S. Shelf	1985	290,270	52,433	-	1,849	-	14
Northeast U.S. Shelf	1986	276,710	42,763	-	2,152	-	14
Northeast U.S. Shelf	1987	271,898	37,210	-	2,270	-	14
Northeast U.S. Shelf	1988	274,747	32,278	-	2,834	-	14
Northeast U.S. Shelf	1989	265,998	32,219	-	3,908	-	14
Northeast U.S. Shelf	1990	276,399	34,031	-	4,576	-	14
Northeast U.S. Shelf	1991	247,255	34,990	-	4,473	-	14
Northeast U.S. Shelf	1992	203,518	33,527	-	4,257	-	14
Northeast U.S. Shelf	1993	160,959	31,056	-	4,476	-	14
Northeast U.S. Shelf	1994	133,506	35,092	-	4,311	-	14
Northeast U.S. Shelf	1995	137,519	33,451	-	4,909	-	14
Northeast U.S. Shelf	1996	150,559	32,780	-	5,543	-	14
Northeast U.S. Shelf	1997	157,528	40,337	-	5,172	-	14
Northeast U.S. Shelf	1998	158,440	43,270	-	4,852	-	14
Northeast U.S. Shelf	1999	160,913	42,567	-	4,046	-	14
Northeast U.S. Shelf	2000	198,862	45,953	-	3,239	-	14
Northeast U.S. Shelf	2001	244,508	45,306	-	2,541	-	14
Northeast U.S. Shelf	2002	237,761	39,697	-	2,053	-	14
Northeast U.S. Shelf	2003	252,053	37,429	-	2,019	-	14
Northeast U.S. Shelf	2004	250,128	30,942	-	2,220	-	14
Northeast U.S. Shelf	2005	252,222	26,540	-	2,605	-	14
Northeast U.S. Shelf	2006	265,773	32,384	-	2,920	-	14

Table S6. Survey biomass estimates and mean maximum length (cm) from all surveys combined and for each of the three focal regions (data plotted in Figs 4E-H, M-P). An asterisk next to the year indicates there were too few surveys conducted in that year for the “All data” values to be meaningful.

Survey name	Year	Demersal	Demersal	Demersal	Invertebrate	Pelagic	Mean	
		≥90	30-90	≤30			Lmax	<i>n</i> surveys
All data	1951*	60.20	20.86	NA	NA	NA	80.7	1
All data	1952*	43.20	19.97	1.04	NA	0.59	98.0	1
All data	1953*	57.32	16.76	1.07	NA	1.16	96.9	2
All data	1954*	74.43	4.38	1.07	NA	0.61	115.5	2
All data	1955*	23.68	2.46	1.30	NA	0.62	127.5	2
All data	1956*	43.59	12.27	1.19	NA	0.56	108.0	2
All data	1957*	29.39	20.84	1.27	NA	0.55	94.8	2
All data	1958*	30.74	20.20	5.04	NA	0.56	97.2	2
All data	1959	24.00	9.53	1.44	1.70	1.20	107.0	4
All data	1960	21.89	14.26	1.81	1.72	0.84	99.2	4
All data	1961	20.45	14.93	1.49	1.59	0.88	98.9	3
All data	1962	19.31	16.37	1.51	1.38	1.55	89.5	4
All data	1963	15.01	13.72	1.57	0.89	2.81	96.0	7
All data	1964	15.95	10.70	1.71	0.87	1.80	92.8	6
All data	1965	15.03	15.81	1.22	0.61	1.83	88.1	6
All data	1966	14.03	13.08	1.40	1.01	1.57	88.4	6
All data	1967	11.92	11.38	1.33	0.76	1.60	92.2	8
All data	1968	10.67	12.69	1.32	0.99	1.97	92.1	8
All data	1969	11.32	9.32	1.38	1.35	1.68	94.2	7
All data	1970	10.10	10.16	1.60	1.32	1.67	86.2	9
All data	1971	6.62	8.53	1.33	1.53	1.44	79.9	9
All data	1972	7.89	8.95	1.60	1.76	1.53	82.3	11
All data	1973	9.60	8.16	1.26	2.17	1.30	85.0	11
All data	1974	6.83	7.76	1.39	2.19	1.53	77.3	10
All data	1975	7.99	8.13	1.42	2.07	1.96	79.5	10
All data	1976	10.40	9.20	1.21	2.67	1.29	79.1	11
All data	1977	8.77	8.66	1.09	1.38	1.12	82.7	12
All data	1978	7.76	7.65	1.15	1.44	1.05	87.6	12
All data	1979	8.81	7.26	0.96	2.05	1.33	91.9	12
All data	1980	8.64	8.26	1.10	2.14	1.45	87.5	15
All data	1981	12.16	9.21	1.38	2.87	1.66	92.8	12
All data	1982	8.76	7.71	1.01	1.90	1.21	89.0	14
All data	1983	12.09	9.09	0.99	1.72	1.58	95.4	14
All data	1984	10.00	8.31	1.09	1.72	1.79	93.6	13
All data	1985	12.69	9.26	1.29	1.59	2.42	93.8	12
All data	1986	11.22	8.81	1.41	1.59	1.86	88.5	15
All data	1987	11.72	9.95	1.36	2.14	3.02	88.6	15
All data	1988	11.00	9.32	1.20	3.08	2.50	88.0	14
All data	1989	9.02	10.76	1.25	3.04	2.38	84.7	15
All data	1990	10.15	10.00	1.33	2.49	2.19	84.7	16
All data	1991	10.43	8.97	1.30	2.61	2.55	87.1	16
All data	1992	8.42	9.12	1.69	2.45	2.88	84.3	15
All data	1993	6.40	8.10	1.52	2.18	3.84	79.8	15

All data	1994	6.65	7.46	1.27	2.62	3.26	82.9	15
All data	1995	7.02	10.34	1.70	2.47	4.05	77.3	16
All data	1996	9.21	9.47	1.63	1.88	2.18	86.4	13
All data	1997	9.59	9.12	1.28	1.62	2.67	86.8	14
All data	1998	8.86	8.41	1.11	1.34	2.69	87.9	13
All data	1999	10.93	8.85	1.28	1.57	3.02	87.6	13
All data	2000	10.37	10.23	1.26	1.65	2.86	84.8	13
All data	2001	11.13	9.88	1.53	2.10	3.16	86.9	14
All data	2002	12.58	10.21	1.48	2.33	3.56	86.1	13
All data	2003	12.01	10.57	1.56	2.05	3.56	81.3	13
All data	2004	12.16	8.22	1.19	2.05	3.22	88.0	13
All data	2005	10.21	8.25	1.34	1.90	2.83	82.9	12
All data	2006	10.54	9.12	1.10	2.94	2.81	86.8	10
All data	2007	10.53	8.75	1.43	2.10	2.91	83.2	9
All data	2008*	7.20	8.96	NA	2.19	4.27	80.8	1
Eastern Bering Sea	1982	88.54	117.18	0.00	52.44	0.20	63.5	1
Eastern Bering Sea	1983	166.71	125.08	0.00	42.57	1.74	71.5	1
Eastern Bering Sea	1984	134.92	119.58	0.00	50.53	0.58	68.5	1
Eastern Bering Sea	1985	126.70	94.39	0.00	26.24	0.78	73.9	1
Eastern Bering Sea	1986	137.42	95.13	0.00	32.83	0.38	74.7	1
Eastern Bering Sea	1987	143.48	116.31	0.00	62.13	0.23	68.6	1
Eastern Bering Sea	1988	185.86	125.62	0.00	70.75	3.36	70.1	1
Eastern Bering Sea	1989	156.88	116.01	0.00	71.87	0.20	67.9	1
Eastern Bering Sea	1990	182.87	117.44	0.00	75.28	0.22	69.2	1
Eastern Bering Sea	1991	133.93	123.41	0.00	78.68	0.93	64.7	1
Eastern Bering Sea	1992	117.94	125.31	0.00	65.46	0.39	65.7	1
Eastern Bering Sea	1993	146.45	141.37	0.00	64.54	3.66	68.6	1
Eastern Bering Sea	1994	150.61	166.26	0.00	64.71	0.90	70.0	1
Eastern Bering Sea	1995	150.86	131.41	0.00	69.07	1.34	69.4	1
Eastern Bering Sea	1996	101.72	139.62	0.00	70.12	0.62	65.7	1
Eastern Bering Sea	1997	90.59	153.41	0.00	87.33	0.97	61.7	1
Eastern Bering Sea	1998	71.56	135.34	0.00	64.90	0.44	62.5	1
Eastern Bering Sea	1999	101.51	95.01	0.00	53.04	0.57	68.9	1
Eastern Bering Sea	2000	132.27	110.04	0.00	66.87	0.83	68.5	1
Eastern Bering Sea	2001	120.40	128.48	0.00	62.83	1.18	68.7	1
Eastern Bering Sea	2002	129.22	119.00	0.00	64.75	0.39	67.5	1
Eastern Bering Sea	2003	200.42	133.08	0.00	67.07	1.17	71.5	1
Eastern Bering Sea	2004	106.54	145.39	0.00	70.08	2.10	64.5	1
Eastern Bering Sea	2005	139.98	159.51	0.00	75.00	2.55	66.5	1
Eastern Bering Sea	2006	86.82	141.53	0.00	70.82	0.59	63.4	1
Eastern Bering Sea	2007	113.60	128.18	0.00	64.77	0.70	66.3	1
Eastern Bering Sea	2008	81.95	128.88	0.00	65.65	1.80	63.1	1
Southern Gulf of St Lawrence	1971	55.63	61.52	0.00	0.00	8.52	121.6	1
Southern Gulf of St Lawrence	1972	63.78	73.05	0.00	0.00	12.16	119.0	1
Southern Gulf of St Lawrence	1973	63.31	74.95	0.00	0.00	15.97	113.3	1
Southern Gulf of St Lawrence	1974	63.76	127.47	0.04	0.00	14.52	102.9	1
Southern Gulf of St Lawrence	1975	42.96	95.34	0.00	0.00	7.23	106.3	1
Southern Gulf of St Lawrence	1976	49.45	141.25	0.00	0.00	2.54	106.3	1
Southern Gulf of St Lawrence	1977	70.35	154.61	0.02	0.00	4.39	109.4	1
Southern Gulf of St Lawrence	1978	127.53	107.72	0.00	0.00	6.61	136.6	1

Southern Gulf of St Lawrence	1979	146.36	141.06	0.00	0.00	1.36	134.1	1
Southern Gulf of St Lawrence	1980	166.62	119.09	0.00	0.00	1.69	143.7	1
Southern Gulf of St Lawrence	1981	251.21	125.42	0.00	0.00	1.50	152.7	1
Southern Gulf of St Lawrence	1982	206.59	72.52	0.01	0.00	1.91	164.2	1
Southern Gulf of St Lawrence	1983	150.67	84.87	0.00	0.00	0.76	150.4	1
Southern Gulf of St Lawrence	1984	123.79	47.04	0.01	0.00	10.41	154.1	1
Southern Gulf of St Lawrence	1985	212.80	61.88	0.01	0.00	20.03	157.8	1
Southern Gulf of St Lawrence	1986	174.56	80.95	0.01	0.00	18.32	145.9	1
Southern Gulf of St Lawrence	1987	132.32	61.62	0.02	0.00	20.63	138.2	1
Southern Gulf of St Lawrence	1988	199.90	98.82	0.02	0.00	13.78	147.2	1
Southern Gulf of St Lawrence	1989	158.67	77.43	0.02	0.00	13.60	143.7	1
Southern Gulf of St Lawrence	1990	118.25	90.34	0.04	0.00	30.64	117.6	1
Southern Gulf of St Lawrence	1991	78.36	75.93	0.06	0.00	37.27	115.4	1
Southern Gulf of St Lawrence	1992	54.36	67.96	0.04	0.00	15.36	105.5	1
Southern Gulf of St Lawrence	1993	72.40	44.62	0.08	0.00	11.42	130.8	1
Southern Gulf of St Lawrence	1994	49.24	41.49	0.08	0.00	15.72	122.8	1
Southern Gulf of St Lawrence	1995	59.35	38.62	0.11	0.00	25.22	122.7	1
Southern Gulf of St Lawrence	1996	61.18	36.41	0.08	0.00	5.15	144.4	1
Southern Gulf of St Lawrence	1997	53.46	27.67	0.11	0.00	17.77	131.3	1
Southern Gulf of St Lawrence	1998	46.22	30.80	0.09	0.00	7.64	137.8	1
Southern Gulf of St Lawrence	1999	61.69	31.76	0.10	0.00	17.88	132.5	1
Southern Gulf of St Lawrence	2000	49.13	36.60	0.07	0.00	12.91	126.7	1
Southern Gulf of St Lawrence	2001	42.19	33.51	0.05	0.00	18.67	121.4	1
Southern Gulf of St Lawrence	2002	73.38	31.24	0.12	0.00	22.94	137.3	1
Southern Gulf of St Lawrence	2003	19.57	40.55	0.09	0.00	34.04	89.9	1
Southern Gulf of St Lawrence	2004	42.65	32.04	0.09	0.00	33.60	113.3	1
Southern Gulf of St Lawrence	2005	20.84	37.59	0.22	0.00	41.88	87.4	1
Southern Gulf of St Lawrence	2006	27.34	34.86	0.13	0.00	16.00	110.3	1
Southern Gulf of St Lawrence	2007	27.00	31.43	0.16	0.00	124.62	72.3	1
Georges Bank	1963	126.86	51.91	0.07	0.02	1.51	107.2	1
Georges Bank	1964	144.93	28.28	0.01	0.35	1.59	109.7	1
Georges Bank	1965	111.44	32.61	0.03	0.86	0.87	105.4	1
Georges Bank	1966	64.16	25.32	0.25	0.06	1.45	105.0	1
Georges Bank	1967	49.20	22.87	0.24	0.63	1.23	110.4	1
Georges Bank	1968	37.96	26.89	0.10	0.71	1.28	102.2	1
Georges Bank	1969	19.01	24.43	0.22	1.27	0.46	92.7	1
Georges Bank	1970	37.89	27.13	0.13	1.53	0.33	101.8	1
Georges Bank	1971	19.65	19.74	0.29	1.53	1.43	96.0	1
Georges Bank	1972	48.18	23.36	0.31	1.35	1.56	119.3	1
Georges Bank	1973	94.09	40.32	0.14	4.83	0.46	121.0	1
Georges Bank	1974	24.79	21.39	0.45	2.37	1.12	96.1	1
Georges Bank	1975	46.42	31.01	0.34	2.71	1.38	112.7	1
Georges Bank	1976	126.31	27.60	0.03	11.58	2.25	119.2	1
Georges Bank	1977	71.73	38.46	0.13	3.90	0.46	103.8	1
Georges Bank	1978	106.70	34.12	0.67	8.39	0.79	122.0	1
Georges Bank	1979	141.93	28.70	0.07	8.07	1.24	124.8	1
Georges Bank	1980	63.44	33.44	0.23	4.73	0.87	105.4	1
Georges Bank	1981	118.55	31.20	0.22	3.40	3.85	124.4	1
Georges Bank	1982	75.16	31.87	0.20	2.82	1.26	114.1	1
Georges Bank	1983	116.65	19.62	0.04	3.08	3.02	123.2	1

Georges Bank	1984	146.96	26.25	0.06	3.61	1.02	127.6	1
Georges Bank	1985	139.56	17.30	0.17	3.47	4.74	127.3	1
Georges Bank	1986	164.11	20.37	0.22	6.09	1.92	124.7	1
Georges Bank	1987	155.61	20.43	0.10	1.48	1.60	132.1	1
Georges Bank	1988	112.97	15.83	0.06	7.73	5.94	122.6	1
Georges Bank	1989	75.12	25.40	0.18	8.98	0.87	113.1	1
Georges Bank	1990	213.32	21.00	0.50	3.85	3.16	138.3	1
Georges Bank	1991	96.57	14.07	0.06	8.16	2.50	127.0	1
Georges Bank	1992	101.35	17.62	0.13	4.52	4.75	129.5	1
Georges Bank	1993	76.39	16.47	0.20	6.19	7.80	120.2	1
Georges Bank	1994	35.11	17.33	0.09	7.39	5.47	102.4	1
Georges Bank	1995	87.50	23.96	0.04	3.30	10.02	118.1	1
Georges Bank	1996	67.63	22.92	0.17	1.27	6.18	116.6	1
Georges Bank	1997	86.94	26.06	0.42	2.42	5.46	122.6	1
Georges Bank	1998	149.64	43.51	0.08	2.33	6.17	126.4	1
Georges Bank	1999	99.20	26.13	0.49	8.79	3.41	117.4	1
Georges Bank	2000	69.82	34.02	0.07	6.68	3.97	103.1	1
Georges Bank	2001	117.70	40.81	0.21	3.05	4.52	118.8	1
Georges Bank	2002	177.88	39.65	0.07	5.77	3.20	119.4	1
Georges Bank	2003	110.91	29.37	0.11	4.32	8.16	120.2	1
Georges Bank	2004	169.92	25.50	0.20	1.70	2.85	123.3	1
Georges Bank	2005	280.75	20.23	0.14	3.57	1.96	141.9	1
Georges Bank	2006	179.32	24.69	0.86	4.42	2.92	132.6	1
Georges Bank	2007	195.69	25.95	0.37	3.71	3.09	131.3	1

Table S7. Number of stock assessments included in the collapse analyses, and the number of these that were “collapsed”, i.e. with biomass less than 20% of B_{MSY} (data plotted in Fig. 4M-P). Where surplus production model fits were used to obtain B_{MSY} , this definition of collapse corresponds to total biomass falling below 10% of pre-exploitation biomass.

Year	All data		Eastern Bering Sea		Eastern Canada		Northeast U.S. Shelf	
	Stocks	Collapsed	Stocks	Collapsed	Stocks	Collapsed	Stocks	Collapsed
1950	24	0	-	-	-	-	-	-
1951	30	1	-	-	-	-	-	-
1952	33	0	-	-	-	-	-	-
1953	33	0	-	-	-	-	-	-
1954	33	0	-	-	-	-	-	-
1955	35	0	-	-	-	-	-	-
1956	38	0	-	-	-	-	-	-
1957	39	0	-	-	-	-	-	-
1958	41	0	-	-	-	-	-	-
1959	47	1	-	-	-	-	-	-
1960	51	2	-	-	-	-	-	-
1961	54	2	-	-	-	-	-	-
1962	54	2	-	-	-	-	-	-
1963	58	2	-	-	-	-	-	-
1964	65	1	-	-	-	-	-	-
1965	65	0	-	-	-	-	-	-
1966	68	1	-	-	-	-	-	-
1967	69	2	-	-	-	-	-	-
1968	73	3	-	-	-	-	-	-
1969	75	3	-	-	-	-	-	-
1970	80	2	-	-	-	-	-	-
1971	83	1	-	-	-	-	-	-
1972	93	1	-	-	-	-	-	-
1973	99	1	-	-	-	-	-	-
1974	102	0	-	-	-	-	-	-
1975	106	2	-	-	6	1	-	-
1976	109	3	-	-	6	1	-	-
1977	121	4	13	1	7	1	-	-
1978	131	3	14	1	8	0	10	1
1979	134	5	15	1	8	0	10	2
1980	139	3	15	1	8	0	11	1
1981	143	1	17	1	8	0	12	0
1982	147	0	17	0	8	0	14	0
1983	149	1	17	0	8	0	14	0
1984	156	5	17	2	8	0	14	1
1985	158	5	18	2	8	0	14	1
1986	163	6	18	2	8	0	14	1
1987	164	6	18	1	8	0	14	2
1988	164	7	18	1	8	0	14	3
1989	165	9	19	2	8	1	14	3
1990	166	5	20	0	8	1	14	3
1991	166	8	20	0	8	2	14	3
1992	166	10	20	0	8	3	14	3

1993	166	14	20	0	8	4	14	5
1994	166	13	20	0	8	4	14	3
1995	166	13	20	0	8	4	14	3
1996	166	10	20	0	8	3	14	3
1997	166	12	20	1	8	3	14	4
1998	166	11	20	1	8	3	14	4
1999	166	13	20	1	8	3	14	3
2000	166	13	20	2	8	3	14	2
2001	165	13	20	1	8	4	14	2
2002	163	14	20	1	8	4	14	2
2003	162	17	20	1	8	5	14	2
2004	159	19	20	1	8	5	11	3
2005	153	19	20	1	6	4	11	3
2006	148	21	20	1	6	4	10	3
2007	81	11	20	1	-	-	-	-

Supporting Figures

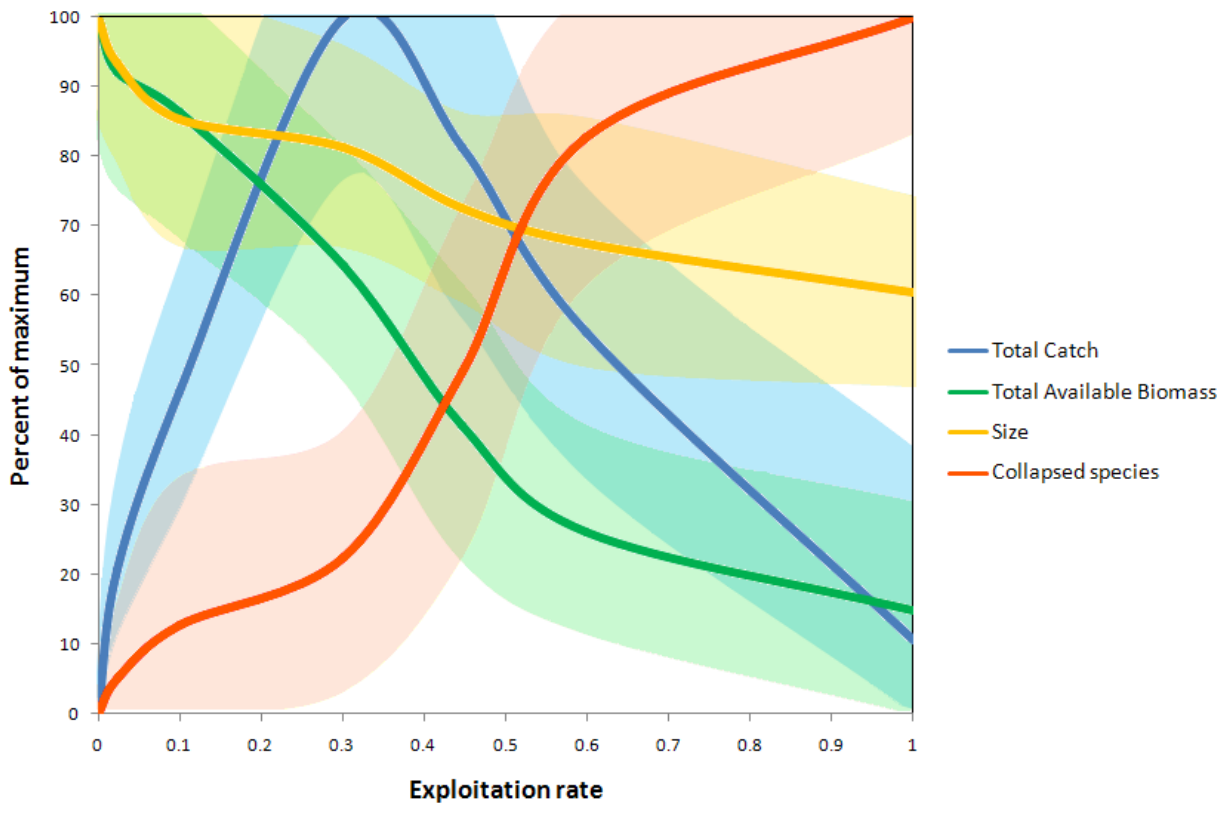


Figure S1: Effects of increasing exploitation rate on 31 model fish communities. Averaged results of ECOSIM models are displayed, shades refer to 95% confidence bounds.

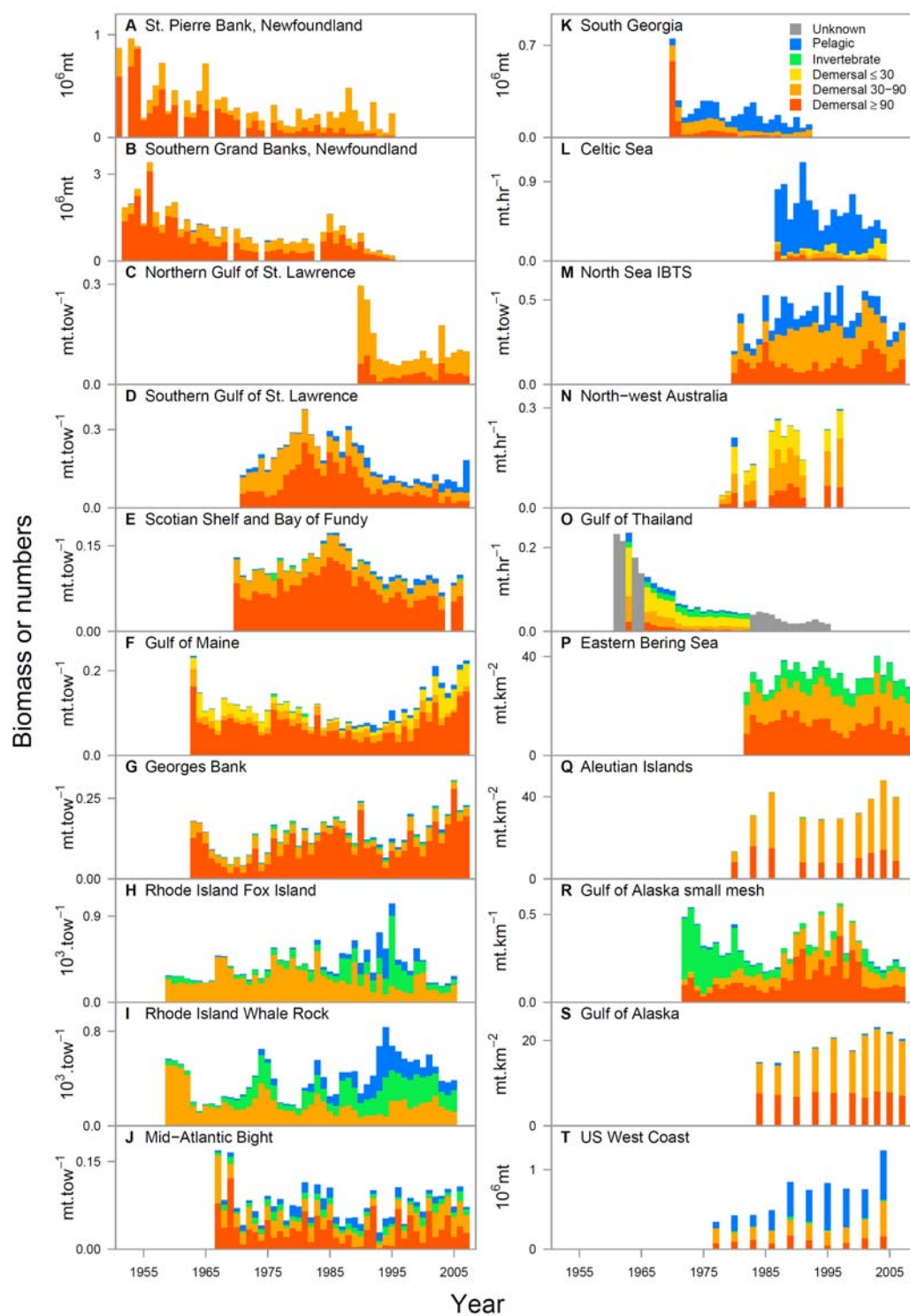


Figure S2. Available survey biomass estimates from 20 ecosystems. Data are grouped into five categories: invertebrate, pelagic (midwater species), and demersal taxa with maximum lengths of ≤ 30 cm, 30-90 cm, and ≥ 90 cm.

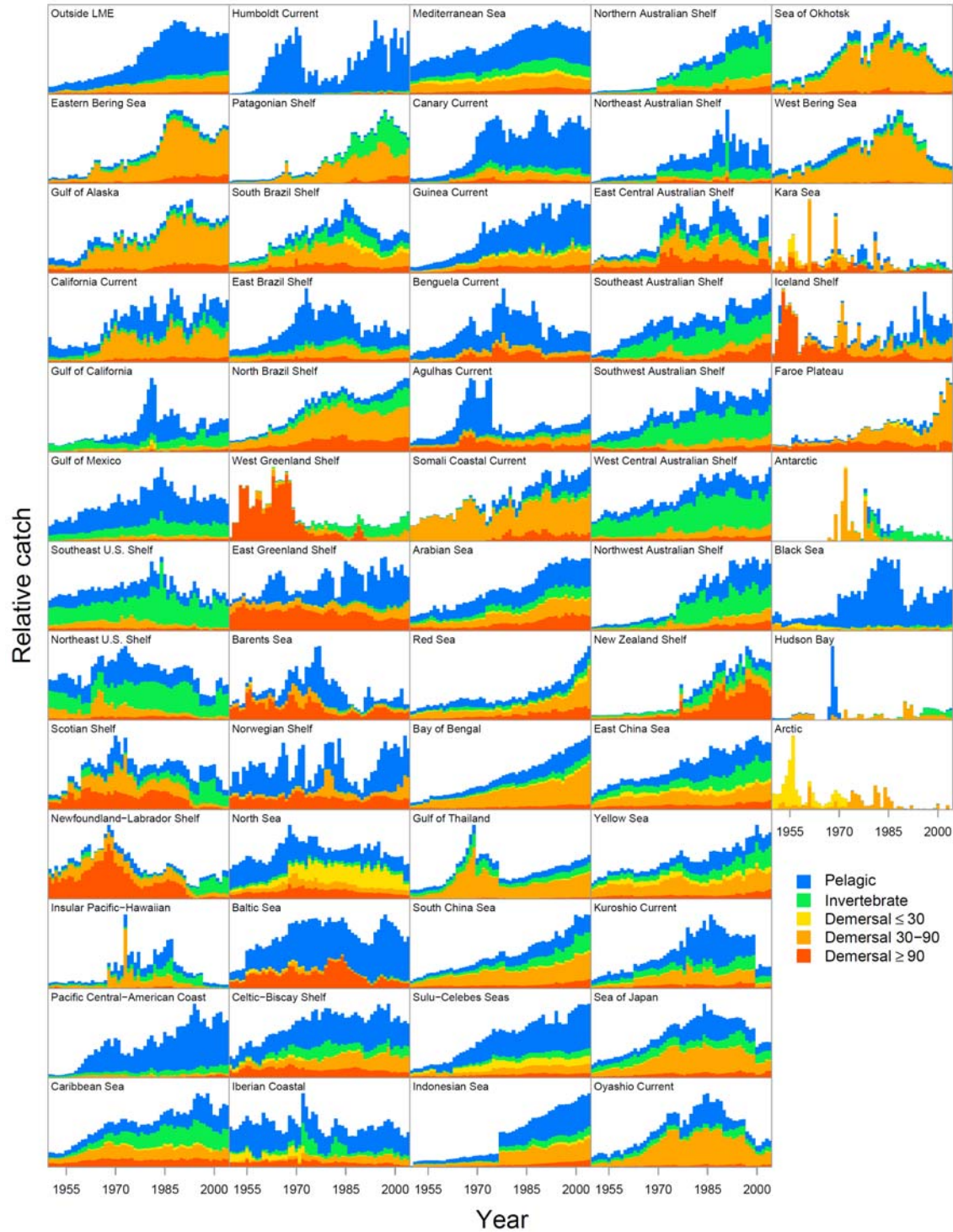


Figure S3: Global catches from the Sea Around Us database, reported by Large Marine Ecosystem and species group.

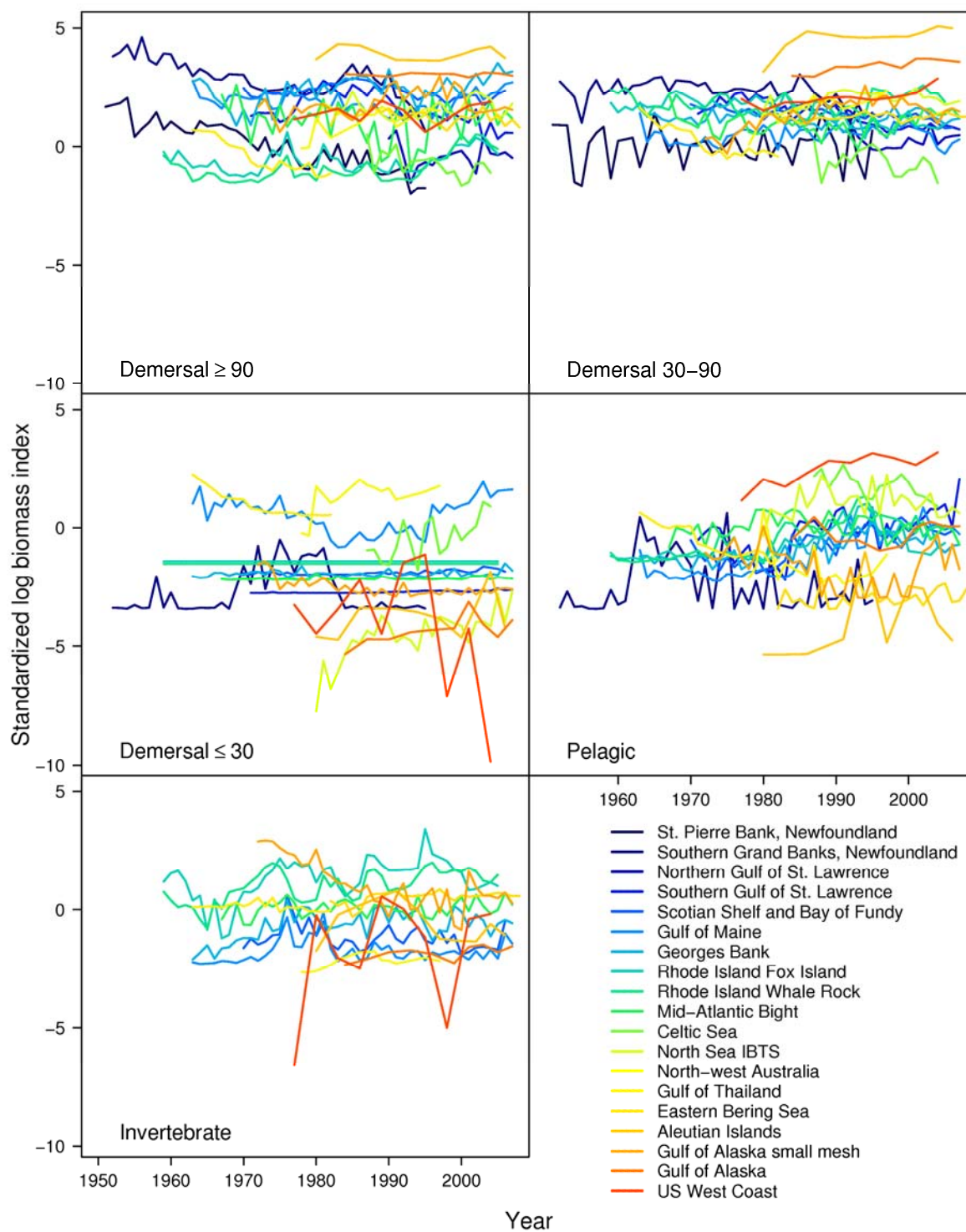


Figure S4: Standardized log biomass indices for each survey and category over time.

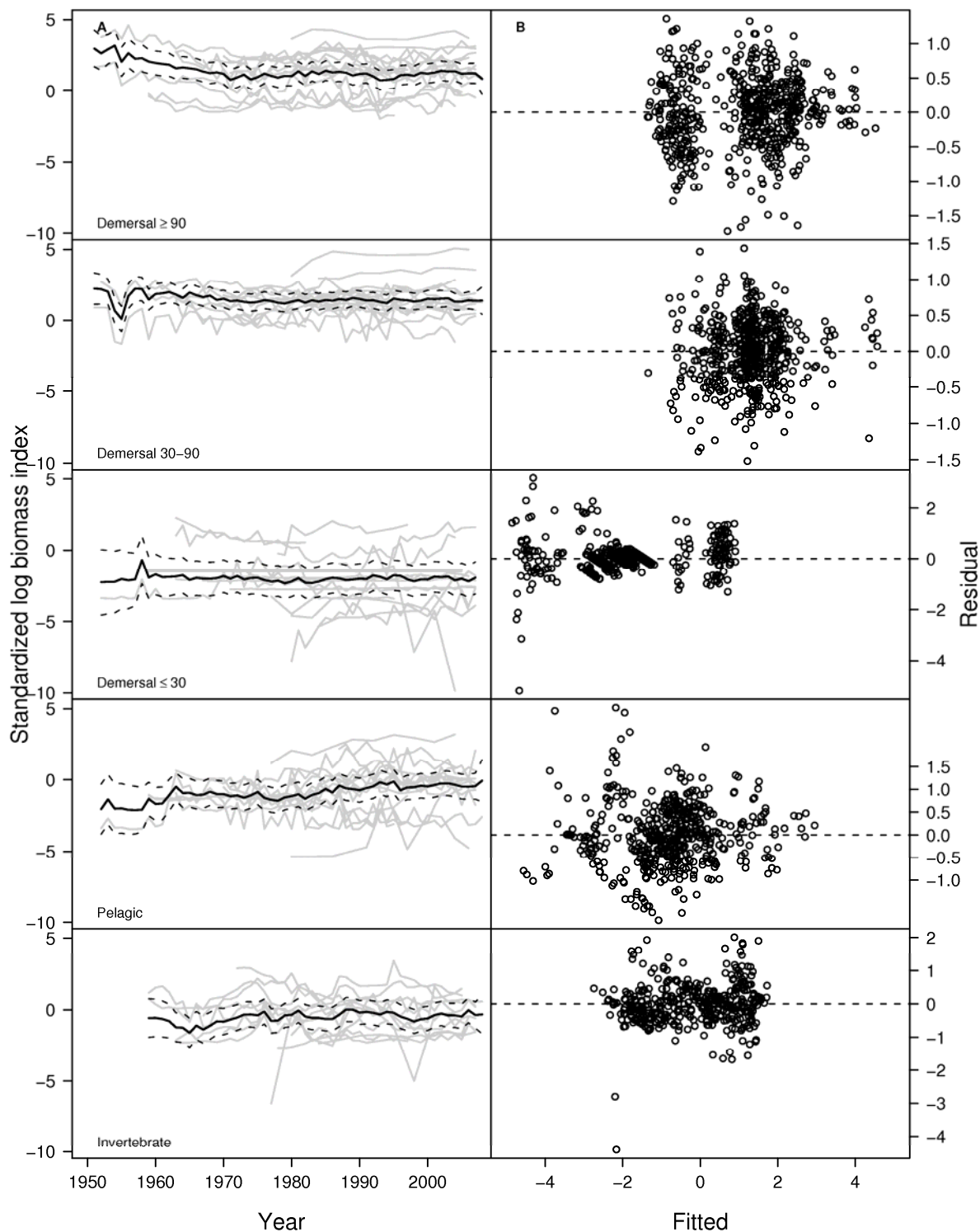


Figure S5: Average trend in survey biomass trends as estimated using a linear mixed effects analysis with a continuous AR(1) within-group correlation structure (Equation 6). Solid and dashed black lines in panel (A) indicate overall fixed effect trends and 95% confidence intervals, respectively; grey lines represent individual survey trends. The fitted values and residuals are plotted in panel (B).

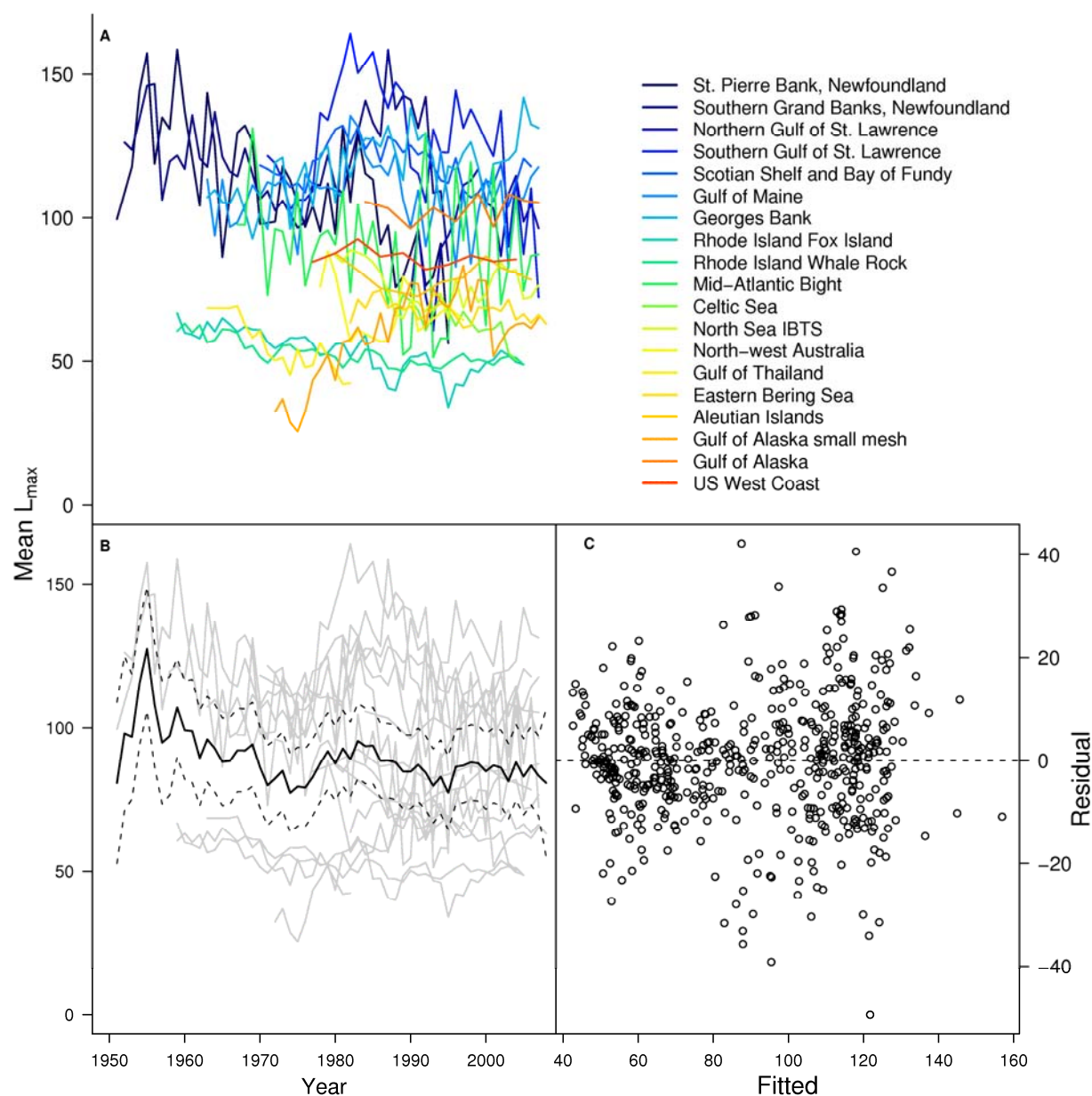


Figure S6. Average trend in L_{\max} , as estimated using a linear mixed effects analysis with a continuous AR(1) within-group correlation structure (Equation 10). Individual survey trends are presented in Panel (A). Solid and dashed black lines in panel (B) indicate overall fixed effects trends and 95% confidence intervals, respectively; grey lines represent individual survey trends.. The fitted values and residuals are plotted in panel (C).

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